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Abstract

Understanding the ways in which tree species interact with solar radiation has previously focused on transmission and reflection of sunlight, typically by examining individual leaves. Here we used a tree crown spectroscopy measurement method to conduct in-situ tests on the radiative performance of ten commonly planted tree species in the UK. Tree crown transflectance (comprehensive effect of transmission and reflection) was examined to determine i), how radiative performance of individual trees varies spatially within a species, and ii), how infrared radiative performance differs between tree species. Our results show that tree crown transflectance depends on the combination of tree crown morphology, local foliage distribution (leaf density, gaps in crown foliage contour, concave or convex crown shapes), solar altitude and leaf size. Spatially, the strongest tree crown transflection was found primarily towards sky on the sunlit side of trees rather than towards the zenith, meaning that infrared transflection towards surrounding buildings and pedestrians is substantial. For all ten species, the tree crown transflectance in the frontal sunlit area was linearly correlated with solar altitude on sunny days. Hence, a solar altitude of 45° was chosen as the benchmark condition for comparing interspecific differences. Interspecific comparison indicated

that interspecific differences in the infrared radiative performance levels were strongly dependent on leaf size when no obvious gaps or concave shapes were present within the tree crowns. Our findings provide insights for understanding radiative interactions between urban trees and surrounding built environment, as well as for tree species selection in urban heat stress mitigation.

Keywords: Urban trees; Infrared radiative performance; Tree crown spectroscopy; Tree species; Urban microclimate

List of symbols

Nomenclature	
$A_{foliage}$	net foliage area of a measuring patch with foliage gaps, m^2
A_{patch}	total area of a measuring patch in viewing vision of the fiber-optic tip, m^2
A_{void}	void area of a measuring patch with foliage gaps, m^2
$D_{measured}$	sampling distance from the fiber-optic tip to a measuring patch on tree crown contours
VA	viewing angle of the spectrometer fiber-optic tip relative to horizontal plane, °
VR	void ratio of a measuring patch with foliage gaps, –
$\overline{x_i}$	the i-th average component value of the statistical mean spectrum, –
Greek symbols	
α	solar altitude, °
λ	wavelength, nm
τR	transflectance (comprehensive effect of transmission and reflection) of tree crown contour or canopy patch, –
$\tau R_{mean,800-900}$	mean transflectance in the near infrared wavelength range of 800–900 nm, –
$\tau R(\lambda)$	spectral transflectance at wavelength λ nm, –
$\tau R_{foliage}$	net transflectance of foliage excluding foliage gaps in a measuring patch, –
τR_{meas_bgd}	background noise signal in sampling transflectance spectrum, –
$\sigma_{A,x}$	type A standard error (uncertainty) of a variable x, unit is the same as x
Abbreviations	
IR	infrared
LAI	leaf area index
PCA	principal component analysis
SAz	solar azimuth (direction)
SVF	sky view factor
UHI	urban heat island
VIS	visible

1 Introduction

Heat waves frequently hit many cities throughout the world and record temperatures are being experienced in many regions in recent years due to the increasing temperatures associated with greenhouse gas emissions [1][2]. It is documented that heat waves have remarkably negative influences on the health of urban inhabitants and contribute significantly to mortality of residents [1][3][4]. Furthermore, anthropogenic global warming is increasing the frequency, duration and intensity of heat waves [5], with the urban heat island (UHI) effect intensifying heat stress under extreme hot climates [6]. In the pursuit of creating benign urban environments to improve human health and well-being, it is imperative to seek effective solutions or strategies for mitigating heat waves and adapting to climate change.

Trees and green spaces enable the provision of better ecosystem services to urban environments [7]. For example, urban trees help to mitigate heat waves through radiative shading [8] and evapotranspiration [9], resulting in lowered air temperatures that help to regulate outdoor thermal comfort [10][11][12]. Gillner *et al.* [13] demonstrated that street trees played a key role in mitigating effects of heat and drought at highly sealed urban sites. Numerical simulations have shown that urban green coverage (trees and grasslands) helped to mitigate human heat stress under different climates [14][15], and the same point was argued in a systematic review of urban greening [16]. Spatial configuration of trees (individual trees versus different types of spacing and/or aggregations) may show significant, but inconsistent results (including

both positive and negative effects) of urban heat mitigation in cities with different climatic conditions [17][18]. Zhou *et al*. [17] presented that percent cover of trees was more important than their spatial configuration in predicting land surface temperature in Baltimore, while the opposite was found in Sacramento. Reasonable urban tree design approaches have a good performance in mitigating daytime and nighttime UHI effects in urban environments [19][20], and an appropriate combination of vegetation and urban geometry can help to mitigate the adverse effects of UHI and provide a better pedestrian thermal comfort [21][22]. More than that, the cooling effects of urban trees or green roofs through radiative shading and evapotranspiration also contribute to building energy savings [12][23] [24] [25]. It is clear that urban forests and trees make important contributions to cities by providing a multitude of benefits [7]. However, tree species have different physiological responses to heat waves and extreme heat events [26][27][28], depending on heat stress adaptability and water availability for the tree species [29], meaning that some trees cope better with high urban temperatures than others [26][30].

To understand urban tree cooling effects and physiological responses, researchers have been focusing on metrics of assessing outdoor thermal comfort (e.g. surface temperature, mean radiant temperature, physiological equivalent temperature), tree physiological indices (e.g. leaf or crown temperature, leaf area index, stomatal conductance, evapotranspiration rate), determinants and quantification of tree cooling capacity, tree shade effects and so on. Surface temperatures of trees and green spaces

are typically 10–20 °C lower than those of sealed ground or built surfaces exposed to sunlight in summer, leading to a significant reduction of mean radiant temperature [13][30][31]. Leuzinger *et al.* [30] reported tree crown temperatures of ten common tree species planted in Central European cities and declared that tree surface temperatures were circa -1 to +4 °C higher than the ambient temperature. They found trees in parks were significantly cooler than those surrounded by sealed ground and small-leaved trees remained cooler than large-leaved trees. Furthermore, surface temperatures of different tree species varied considerably, and the SVF (sky view factor – the ratio of the amount of the sky that can be seen from a given point on a surface to that potentially available, ranging from 0 to 1) value had a significant effect on tree surface temperatures [18][32]. It is presumed that the mean radiant temperature, which is linked to global temperature, air temperature and wind speed [19], is closely related to the urban thermal comfort [22]. Park *et al.* [33] proposed a multilayer mean radiant temperature model for pedestrians in a street canyon with trees. Physiological equivalent temperature is also an important assessing index of outdoor thermal comfort [18][34]. Zölch *et al.* [35] showed that planting trees had the strongest impact with an average physiological equivalent temperature reduction of 13% compared with existing vegetation. Zheng *et al.* [31] measured the influence of trees on the outdoor thermal environment in subtropical areas through field tests of relevant physiological indices and microclimatic parameters. They found that the widely planted fig tree *Ficus microcarpa* had the best cooling performance among four tree species studied, with the maximum reduction of physiological equivalent temperature due to the highest leaf area

index (LAI).

As to the determinants of tree cooling effects in terms of tree physiological indices, Morakinyo *et al.* [36] revealed that LAI was the main driver of tree cooling for outdoor temperature regulation, followed by trunk height, tree height and crown diameter. LAI was also highlighted by Armson *et al.* [37] and Rahman *et al.* [38]. Zhang *et al.* [39] stated that tall trees with a large LAI and canopy diameter should be a priority to improve the comfort of outdoor environments. Furthermore, different tree species may differ in microclimate benefits. Sanusi *et al.* [40] declared that the microclimatic benefits in streets with *Ulmus procera* and *Platanus* x *acerifolia* trees were significantly greater than the street with *Eucalyptus scoparia* trees, in terms of air temperature, relative humidity, solar radiation, mean radiant temperature, wind speed. In assessing a tree's cooling capacity via transpiration, it is presumed that different tree species have significant difference in evaporative cooling [24][41][42][43]. For example, as reported in [44], *Tilia cordata* trees with higher LAI and sap-wood area provided three times more transpiration than *Robinia pseudoacacia*. Konarska *et al.* [45] observed that night-time transpiration in all the seven species they studied amounted to 7 and 20 % of midday transpiration of sunlit and shaded leaves, respectively, in a high latitude city in Gothenburg, Sweden. There are some other ways of quantifying the cooling effects of urban trees through transpiration. For instance, Wang *et al.* [46] quantified the cooling capacity of urban trees as the surface cooling rate (the negative ratio of land surface temperature changes to fractional tree cover changes). They found that the surface

cooling rate was dominated by plant transpiration, up to 1.336 °C per percentage of fractional tree cover in heat waves in cities of the contiguous United States. Additionally, tree shade provides a good outdoor thermal comfort for pedestrians and enables energy savings. Rahman *et al.* [44][47] investigated vertical air temperature gradients under tree shades during summer days. Tree radiative shading effect was simulated by Upreti *et al.* [8] in a regional built environment, who predicted the capacity of urban trees in reducing urban surface and air temperature by about 2–9 °C and 1–5 °C, respectively. It is argued that shade trees have more prominent energy saving potential than urban lawns in a desert city [12]. Urban lawns tend to be hugely wasteful of water resources, so well chosen species of shade trees or xerophytes may be far more energy efficient than lawns in arid or semi-arid environments [48][49].

Apart from research into urban tree cooling effect, reasonable tree planting strategies were widely considered. Reasonable arrangement of spacing and size of street trees was proven to be beneficial for decreasing pedestrian mean radiant temperature [50]. Different tree planting strategies are available in [39][51][52][53][54][55][56] for improving outdoor thermal comfort in different scenarios and climates.

Generally, previous research into the cooling effect of trees in regulating urban microclimates simplified the physical characterization of radiative performance of trees [8][39][57][58][59]. Some studies have excluded the radiative shading effect of trees in urban microclimatic modelling [60][61]. Where foliage albedo values for different tree

160 species have been incorporated in microclimate modelling to illustrate their influences
161 on thermal environment they are commonly fixed [59], yet the foliage albedo of trees
162 varies temporally and spatially throughout a day (as presented in sections 4.2.3 and 4.3
163 of the present study). To better understand the cooling effect and capacity of urban trees,
164 it is essential to characterise the radiative performance of different tree species from the
165 perspective of physical characteristics, especially in the near infrared (NIR) region. In
166 an earlier study, we established a novel methodology of characterising infrared (IR)
167 radiative performance of urban trees using tree crown spectroscopy [62]. Experimental
168 tests on *Tilia cordata* (aka small-leaved lime or little-leaf linden), a commonly planted
169 tree species in the UK and Europe, were carried out to demonstrate the impact factors
170 of IR radiative performance qualitatively in terms of transflectance at the tree crown
171 level. However, the way in which IR radiative performance varies spatially across a tree
172 crown and interspecific differences remain unanswered. Hence, the present study aims
173 to explore spatial distribution of IR radiative performance across a tree crown, as well
174 as to ascertain differences of IR radiative performance levels in tree species that are
175 often planted in the UK. Since the radiative performance of tree species in terms of the
176 tree crown transflectance varies with solar time throughout a day in a clear sky, as stated
177 in our previous work [62], the way in which transflectance varies with solar time was
178 explored as well, helping to make a benchmark for comparing interspecific
179 performance differences. Ten commonly planted tree species in the UK were chosen
180 for extensive in-situ tests of tree crown transflectance on sunny days using the
181 established spectroscopy measurement method, in order to assess interspecific

182 differences of IR radiative performance levels on the same benchmark.

183

184 ## 2 Test site, method and conditions

185 ### 2.1 Test site and selection of tree species

186 Given the convenience and accessibility of trees for extensive in-situ tests, the test site
187 was chosen at the Whiteknights campus, University of Reading (51.44° N, 0.94° W),
188 UK. The campus covers an area of 123 hectares and features high plant diversity, with
189 an estimated number of 150 different tree species. Ten commonly planted tree species
190 in urban spaces throughout the UK were chosen for tests, as shown in Figure 1. Five of
191 these are native British trees [63]: *Carpinus betulus* (hornbeam), *Acer campestre* (field
192 maple), *Quercus robur* (English oak), *Tilia platyphyllos* (large-leaved lime), *Betula*
193 *pendula* (silver birch).

194

195

196 **Figure 1.** Ten tree species selected in field tests, top row left to right: *Sequoiadendron*

giganteum (giant sequoia), *Carpinus betulus* ('Fastigiata' cultivar, hornbeam), *Acer campestre* (field maple), *Quercus robur* (English oak) and *Platanus* x *acerifolia* (London plane); bottom row left to right: *Tilia platyphyllos* (large-leaved lime), *Acer* x *freemanii* (autumn blaze maple), *Betula pendula* (silver birch), *Acer platanoides* ('Schwedleri' cultivar, copper Norway maple) and *Aesculus hippocastanum* (horse chestnut).

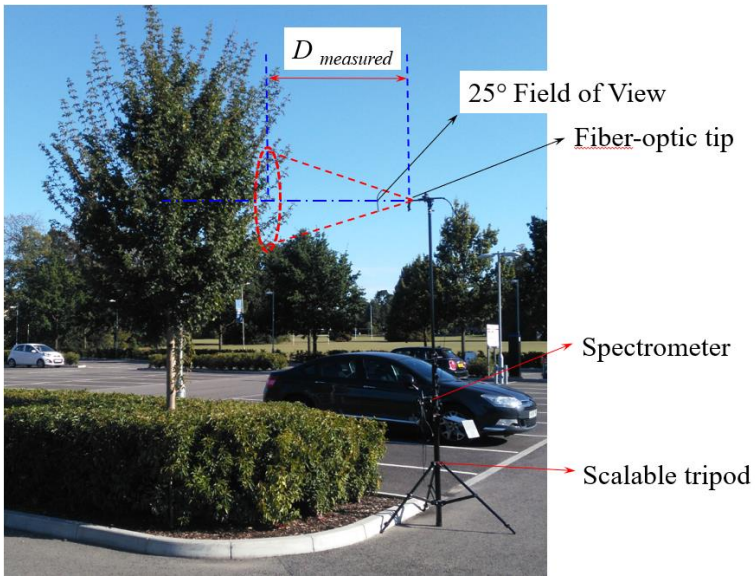
2.2 Test methods, instruments and facilities

We previously established a methodology for characterising IR radiative performance of urban trees using tree crown spectroscopy [62]. The term tree crown ***transflectance*** (τR) or ***transflection***, which represents comprehensive radiative performance of trees at the crown level, is introduced in contrast to reflectance or transmittance at tree leaf levels. This is because when an optical sensor (i.e. spectrometer fiber-optic cable) is positioned at one side of trees to measure the radiative performance of 'a patch of tree crown surfaces' (abbreviated as 'measuring patch' hereafter), light received by the spectrometer usually comprises single-reflected, multi-reflected, multi-transmitted and transmitted-reflected rays through leaves.

A Black-Comet-SR model CXR-SR (StellarNET Inc., Tampa, Florida, USA) concave grating miniature spectrometer with a wavelength range of 350–1000 nm and a spectroradiometer mode was used for in-situ tests of tree crown transflectance. It covers the full visible (VIS) region (350–700 nm) and an important spectrum transition from

219 VIS to NIR around 700 nm where the leaf reflectance and crown transflectance jump
220 sharply (see testing results in section 4). Its spectral resolution is 0.5 nm. In the
221 spectroradiometer mode, the fiber-optic tip can be fitted with a cosine receptor which
222 allows measurement of solar spectra irradiance in a 180° field of view in the 350–1000
223 nm wavelength range. The spectroradiometer mode was only used for solar spectral
224 irradiance measurements in conditions of different viewing angles of the fiber-optic tip.
225
226 Regarding the test facilities, a scalable tripod with a full height of 8 m was used to hold
227 and position the fiber-optic cable of the miniature spectrometer in the field tests, as
228 shown in Figure 2(a). The fiber-optic cable was mounted onto the top of the tripod at
229 one end and connected to a data acquisition laptop at the other. A USB camera was
230 fixed close to the fiber-optic tip in order to capture viewing vision of the latter in field
231 tests. The miniature spectrometer was powered through a USB cable connected to the
232 laptop. The battery when fully charged usually lasted for about five hours powering
233 both the laptop and the spectrometer. The viewing angle of the optical sensor can be
234 adjusted in all directions. The fiber-optic cable was usually used without any cosine
235 receptor and had a field of view of 25°. A reflectance standard RS50 shown in Figure
236 2(b) was adopted to measure reference spectra. The reference spectrum was always
237 sampled in a vertical plane in the transient solar azimuth (SAz) direction, as justified in
238 our previous work [62]. Particularly, for a test sequence of transflectance spatial
239 distribution, a single vertical reference plane was employed, and the test sets were
240 performed within several minutes. Appropriate measuring distance from the trees
241 should be chosen to make the measuring results robust. It was found that the sampling

distance ($D_{measured}$ in Figure 2(a)) from the fiber-optic tip to the measuring patches on the tree crown surfaces could not be too close (e.g. less than 1.5 m), as it would lead to higher transflectance levels than a sampling distance beyond 2.0 m where the measuring result was robust. This measurement issue was greatest for tree species with large-sized leaves such as *Platanus x acerifolia* (London plane), *Aesculus hippocastanum* (horse chestnut), in which the IR transflectance would be overestimated by 25% at a closer sampling distance (e.g. 1.0 m) due to a specular reflectance effect from partial leaf surfaces. In view of this, the sampling distance was kept between 2.0–5.0 m for most of the tested trees, depending on the tree crown diameters and the viewing vision of the USB camera. In some small trees, e.g. *Acer campestre* (field maple) a distance of 1.5–2.0 m was chosen because the height of the trees was less than 5 m and the diameter of the tree crowns was less than 3 m.



(a)

(b) 

Figure 2. (a) A scalable tripod holding the optical fiber spectrometer for field tests; (b) Optic fiber tip of the spectrometer and a reflectance standard RS50.

A model SM2500 spectrometer (Spectral Evolution, Haverhill, Massachusetts, USA) with spectral resolution of 3.5–22 nm in the full range of UV (ultraviolet), VIS, NIR (wavelength range: 350–2500 nm and wavelength reproducibility of 0.1 nm at an accuracy of 0.5 bandwidth) was used in the laboratory to measure leaf reflectance of different tree species, in contrast to the tree crown transflectance levels. The spectrometer was deployed together with a leaf clamp supplied by the manufacturer for leaf reflectance measurements. This spectrometer with a broad spectral range was bulky and expensive, thus not suitable for mounting on a tripod for field tests.

2.3 Test conditions

To explore spatial distribution rules of the radiative performance of individual trees, one fastigiate hornbeam tree of 7.0 m height and one *Acer campestre* tree of 4.4 m height were primarily measured at multiple times, with different viewing angles of the spectrometer fiber-optic tip pointing at different spatial locations of their crowns. Wider field tests of the ten tree species were implemented to explore interspecific differences

of the radiative performance in terms of the transflectance in the sunlit area of trees. At least five individual trees were sampled for each species in order to assess both intraspecific (within-species) and interspecific (between-species) variability in transflectance. All field tests of tree crown transflectance were carried out on sunny days or in sunny time slots without clouds shading the sun, in order to maintain a constant solar irradiance for the reference spectrum and the sampling transflectance spectrum. Measuring heights of the transflectance were usually chosen around the center height of the tree crown within ±1.0 m deviation. Background noise signal was measured for several individual trees to assess its impact on measurements, which was found to be negligible as shown in section 4.2.1.

In addition to meteorological microclimate conditions (outdoor air temperature, incoming and outgoing shortwave radiation, incoming and outgoing longwave radiation), the soil moisture contents and physiological conditions (leaf temperature) of a single fastigiate hornbeam (*Carpinus betulus*) tree were monitored during the testing period of June to September 2019, providing information reference for the test conditions. Chlorophyll fluorescence and heat stress of five tree species were measured to inform their physiological stress status. Reflectance spectra of multiple individual leaves of various tree species were also measured in the lab for contrast.

3 Data processing and error analysis

Repeated measurements of leaf reflectance or tree crown transflectance spectra for

298 various tree species were implemented to obtain the statistical mean. Notate *M* as the
299 number of spectra samples of a specific tree species. For each spectrum sample, assume
300 that *N* rows of spectral values are recorded at different wavelength intervals. The
301 statistical mean spectrum is calculated based on the sample component values in each
302 row, as given in Equation (1).

303 $$\bar{x_i} = \sum_{j=1}^{M} x_{i,j} \quad (1)$$

304 where $\bar{x_i}$ is the *i*-th average component value of the statistical mean spectrum, while
305 $x_{i,j}$ (i = 1, 2, 3,…, N; j = 1, 2, 3,…, M) denotes the *i*-th measured component value
306 in the *j*-th sample spectrum.

307

308 With respect to error analysis of the statistical mean spectrum of the tree crown
309 transflectance or the reflectance of individual leaves, A-type standard error (uncertainty)
310 is usually used to estimate the statistical mean errors [64], as described in Equation (2).

311

312 $$\sigma_{A,x} = \sqrt{\frac{1}{M(M-1)} \sum_{j=1}^{M} \left(x_j - \bar{x}\right)^2} \quad (2)$$

313

314 where $\sigma_{A,x}$ is the A type standard error of a variable x, M is the number of samples,
315 $\bar{x}$ is the statistical mean value of variable x, and x_j represents the *j*-th sample value.

316

317 The standard error of the statistical mean tree crown transflectance or leaf reflectance
318 spectrum at each wavelength band is therefore given by:

319 $$\sigma_{A,\tau R}(i) = \sqrt{\frac{1}{M(M-1)} \sum_{j=1}^{M} \left(x_{i,j} - \bar{x_i}\right)^2} \quad (3)$$

where $\sigma_{A,\tau R}(i)$ is the *i*-th component of the A-type standard error for the transflectance (τR) spectrum.

When calculating the statistical mean, if a spectrum in the samples was found to be outside of the 99.8% confidence interval, i.e. $[-3\sigma_{A,TR}, +3\sigma_{A,TR}]$, it was identified as a spectrum outlier. The statistical mean of the targeted spectrum was then recalculated excluding any outliers to minimize the contribution of measurement errors or non-target biological processes such as damaged or discolored leaves.

3.1 Statistical analyses using PCA (principal component analysis)

Unless otherwise noted all statistical analyses using PCA were carried out in R version 3.6.0 [65]. Packages "vegan" [66], "factoextra" [67] and "FactoMineR" [68] were used to analyse spectral data. Due to the nature of this data, ordination using Principal Component Analysis (PCA) and Principal Coordinates Analysis (PCoA or MDS - Multidimensional Scaling) generate the same outcome; for consistency these ordinations are referred to as PCA. Briefly, PCA takes a multivariate dataset of potentially correlated variables and transforms them into fewer, uncorrelated variables (principal components). This approach is commonly applied in ecology (see Legendre and Legendre [69]) and has previously been applied to leaf spectral data in remote sensing studies (e.g. Cavender-Bares et al. [70]). Spectral data (transflectance measured across the range 350-1000 nm) were interpolated for each of 67 individual trees from

0.5 nm bands into 5 nm bands and then scaled prior to analysis. The 10-fold reduction in data points following interpolation (1300 to 130 bands per tree) led to a loss of only 0.1% of overall explained variance. Significant correlations between PCA axes and spectra were used to assess the specific differences represented by the axes, and species was used as a grouping variable.

Further examination of inter- and intra-specific differences in transflectance profiles for 350-1000 nm was conducted by converting PCA eigenvalues into pairwise euclidean distances and then applying permutational multivariate analysis of variance (PERMANOVA) on the resulting distance matrices. The multivariate homogeneity of dispersion (variance) was then assessed to examine the extent to which observed differences i) could be attributed to interspecific differences in transflectance values, or ii) may have been confounded by differences in intraspecific variance.

4 Results and discussion

Figure 3 illustrate the flow chart of research framework in the discussion, in order to determine i), how radiative performance of individual trees varies spatially within a species, and ii), how infrared radiative performance differs between tree species.

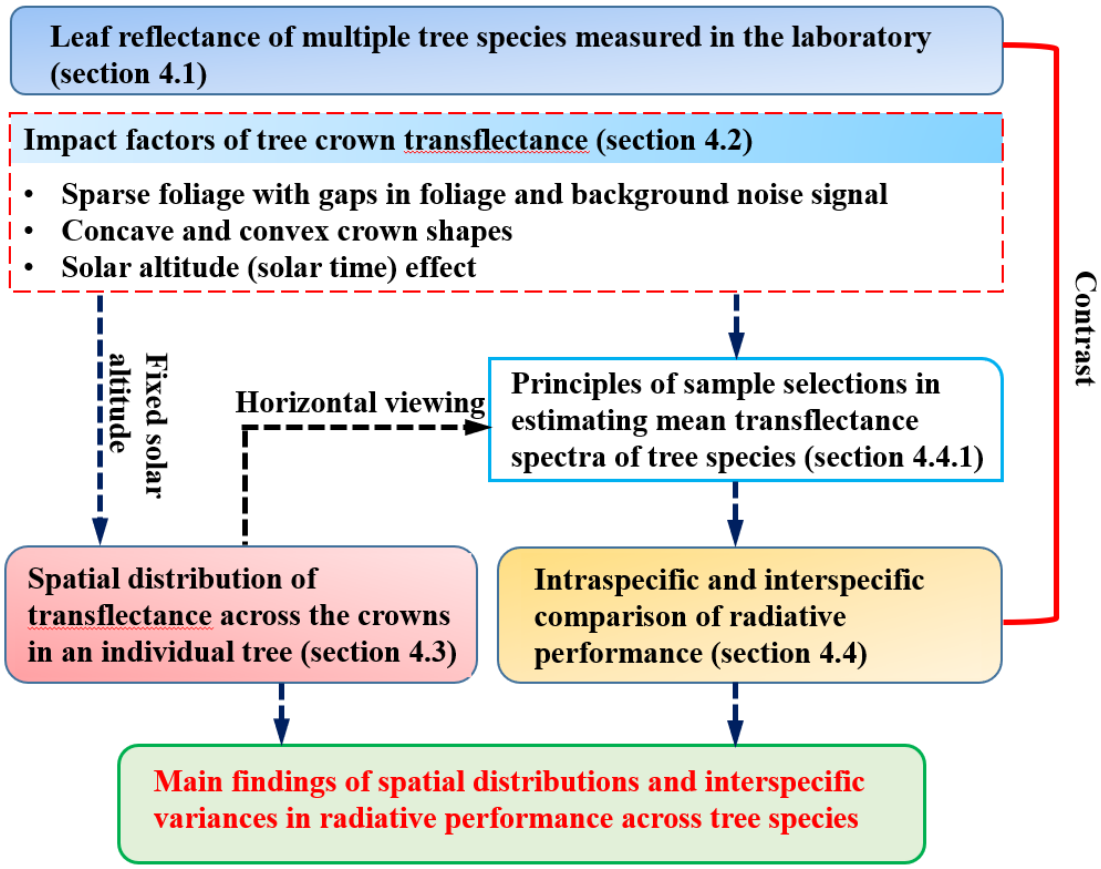
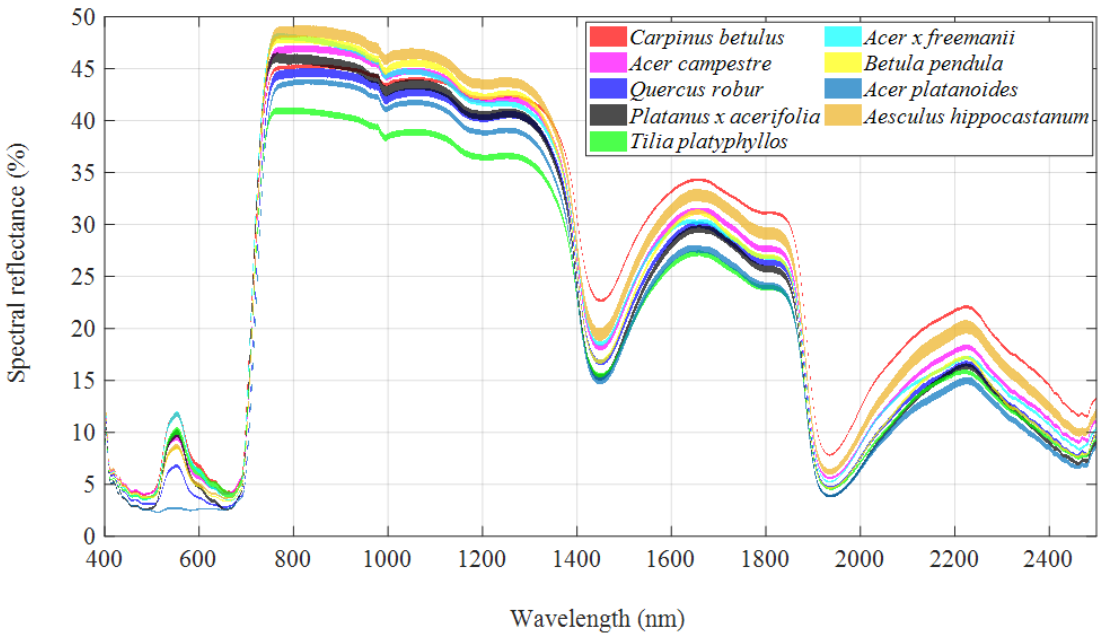


Figure 3. Flow chart of the research framework in discussion.

4.1 Leaf reflectance measured in the laboratory

Leaf reflectance spectra of 9 targeted tree species (except *Sequoiadendron giganteum*) were measured at multiple times in the laboratory using the SM2500 spectrometer (Spectral Evolution) with a leaf clamp, in order to estimate statistical mean leaf reflectance spectra. Leaf reflectance of the species *Sequoiadendron giganteum* was not measured, as the spectrometer could not be used to measure individual needle leaves. For each tree species, 10 leaves were collected randomly from individual trees and 5–10 trees were sampled. Reflectance spectra of the collected leaves were measured within 1 hour after leaf collections to guarantee that the leaves would not lose moisture.

373 Each selected leaf was measured once, and the measuring point was chosen randomly
374 with consideration of avoiding main leaf veins. Figure 4 gives the statistical mean leaf
375 reflectance spectra of 9 tree species, indicating there is a minor difference of mean leaf
376 reflectance spectra among various tree species (within ± 4% of spectral reflectance
377 values). The standard mean errors of spectral reflectance of various tree species in the
378 wavelength range of 400–2500 nm were within ±0.5% with a large number of samples
379 (50–100 samples for each species). Data on statistical mean reflectance spectra of
380 leaves across species can be accessed in ***Appendix A***.
381



382

383 **Figure 4.** Standard mean error bands of leaf reflectance spectra of nine tree species
384 measured in the laboratory and sampled from 5–10 trees for each species.
385

386 **4.2 Impact factors of tree crown transflectance**

387 Through a multitude of in-situ tests, it was found that visibly non-uniform foliage

388 distribution in the measuring patches of tree crown contours, such as sparse foliage with
389 gaps in crown foliage and/or concave crown shapes, would affect tree crown
390 transflectance (τR) to different extents. On the other aspect, our previous work
391 indicated that solar time was one of the most important factors impacting τR levels
392 [62]. It is therefore necessary to elucidate the impacts and determine the principles of
393 sample selections for estimating the radiative performance levels of multiple tree
394 species.
395
396 **4.2.1 Impact of sparse foliage with gaps in crown foliage and background noise**
397 **signal**
398 Sparse foliage is commonly visible at some locations of tree crowns. To estimate
399 measurement errors of the measuring patches with gaps in the sparse foliage and
400 examine the impact of background noise signal penetrating the gaps, it is essential to
401 measure the transflectance of the measuring patch along with the background noise
402 signal. Void ratio of the gaps in the measuring patch should be also determined, in order
403 to correct the net transflectance of the sparse foliage by excluding background noise
404 signal. Total area of a measuring patch with gaps (A_{patch}) is sum of the net foliage area
405 ($A_{foliage}$) and the void area (A_{void}) in equation (5). Void ratio of the gaps (VR) is
406 calculated in equation (6). Relation between the measured transflectance
407 (τR_{meas_patch}), the net transflectance of foliage ($\tau R_{foliage}$) and the background noise
408 signal (τR_{meas_bgd}) is described by equation (7) based on radiative energy conservation.
409 Thus, the net transflectance of foliage ($\tau R_{foliage}$) is rearranged in equation (8) and the
410 relative error is estimated by equation (9).

$$A_{patch} = A_{foliage} + A_{void} \quad (5)$$

$$VR = A_{void}/A_{patch} = A_{void}/\left(A_{foliage} + A_{void}\right) \quad (6)$$

$$\tau R_{meas\_patch} = \frac{\tau R_{foliage} \cdot IRR_{ref} \cdot A_{foliage} + \tau R_{meas\_bgd} \cdot IRR_{ref} \cdot A_{void}}{A_{patch} \cdot IRR_{ref}}$$

$$= \tau R_{foliage} \cdot VR + \tau R_{meas\_bgd} \cdot (1 - VR) \quad (7)$$

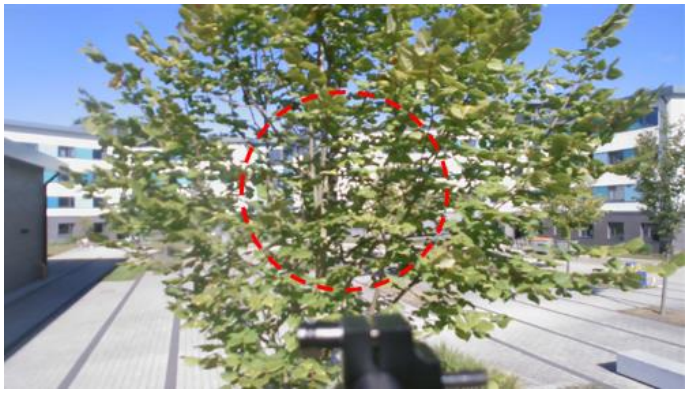
$$\tau R_{foliage} = \tau R_{meas\_patch} + \frac{VR}{1-VR} \cdot \left(\tau R_{meas\_patch} - \tau R_{meas\_bgd}\right) \quad (8)$$

$$Error_{sparsefoliage} = \frac{VR}{1-VR} \cdot \left(1 - \frac{\tau R_{meas\_bgd}}{\tau R_{meas\_patch}}\right) \times 100\% \quad (9)$$

Take the τR measurement of a *Tilia platyphyllos* tree with sparse foliage as an example to illustrate the impact of background noise. Figure 5(a) shows the viewing vision of the tree crown τR measurement in the sunlit area and SAz (solar azimuth) direction with solar altitude α = 34°. The void ratio of gaps (VR) in the vision was estimated by the ImageJ software [71], resulting in VR = 23.3% (see Figure 5(b)). The net τR of the foliage ($\tau R_{foliage}$) was calculated in equation (8). Figure 5(c) shows the corrected net foliage τR spectrum excluding gaps in foliage compared to the measured τR spectra. It suggested that the measurement error by the background noise was very small (within 3% deviation). Furthermore, the net τR of the foliage was lower (-4% deviation of IR τR in this case) than the τR of a dense measuring patch in another *Tilia platyphyllos* tree, indicating that a sparse foliage (sparse leaf density) degraded the τR levels.

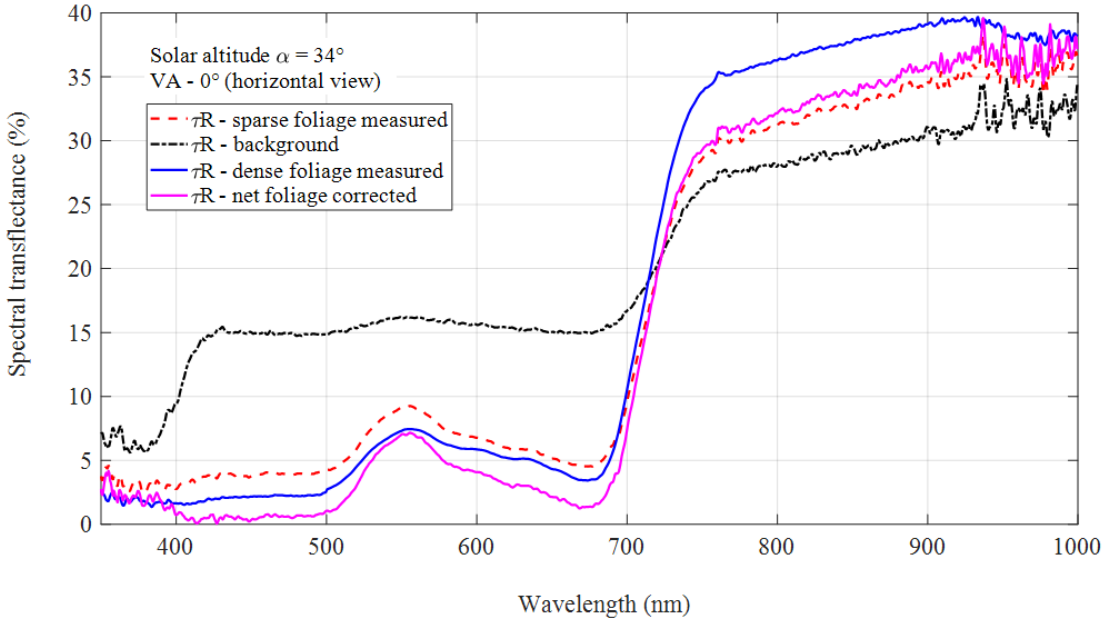
Distributions of sparse foliage on individual trees are amorphous and heterogeneous, resulting in different values of void ratio. When choosing measuring patches with

visibly dense foliage (no obvious gaps, VR < 10%), it is easy to control the measuring errors. For most of the trees in tests, $VR \approx 5\%$, while the observed spectral transflectance between the measuring patches and the background noise signal is usually less than 30% ($\tau R_{meas\_patch} - \tau R_{meas_{bgd}} \leq 30\%$), it implies that the absolute measurement error of the transflectance is below 3% using equation (9). In this sense, background noise in the transflectance measurement with visibly dense foliage can be disregarded.



(a) Viewing vision (circle in red dashed line) of the tree crown transflectance measurement in the SAz direction

(b) Void ratio of gaps in the viewing vision determined by the ImageJ software



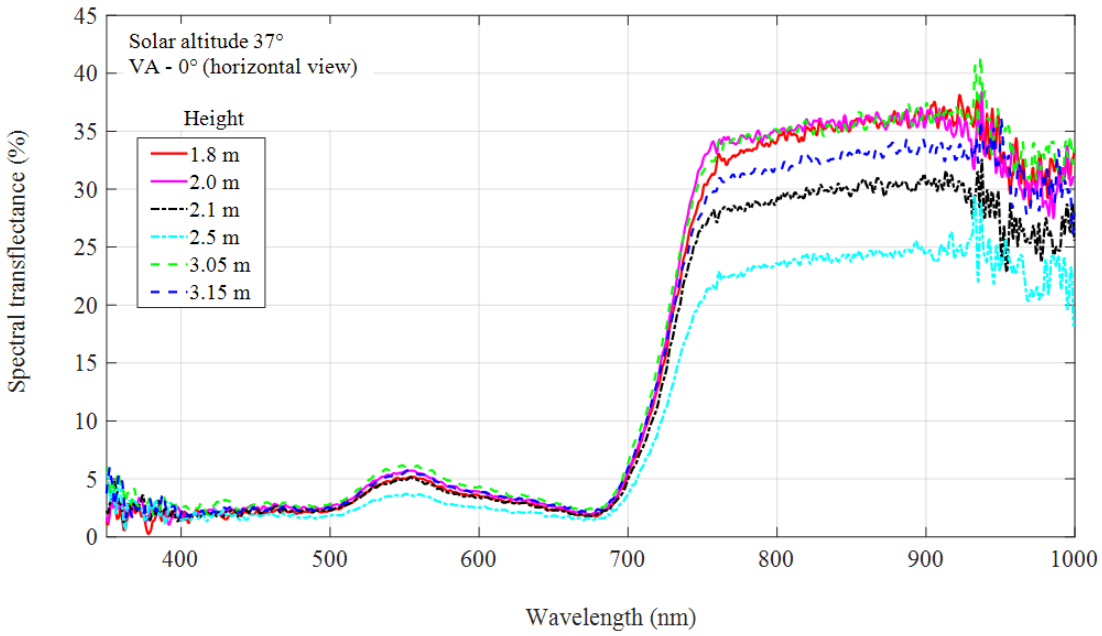
(c) Net transflectance of foliage excluding gaps compared to measured transflectance

Figure 5. Tree crown transflectance of a measuring patch with sparse foliage and gaps in a *Tilia platyphyllos* and estimation of the background noise impact with α = 34°.

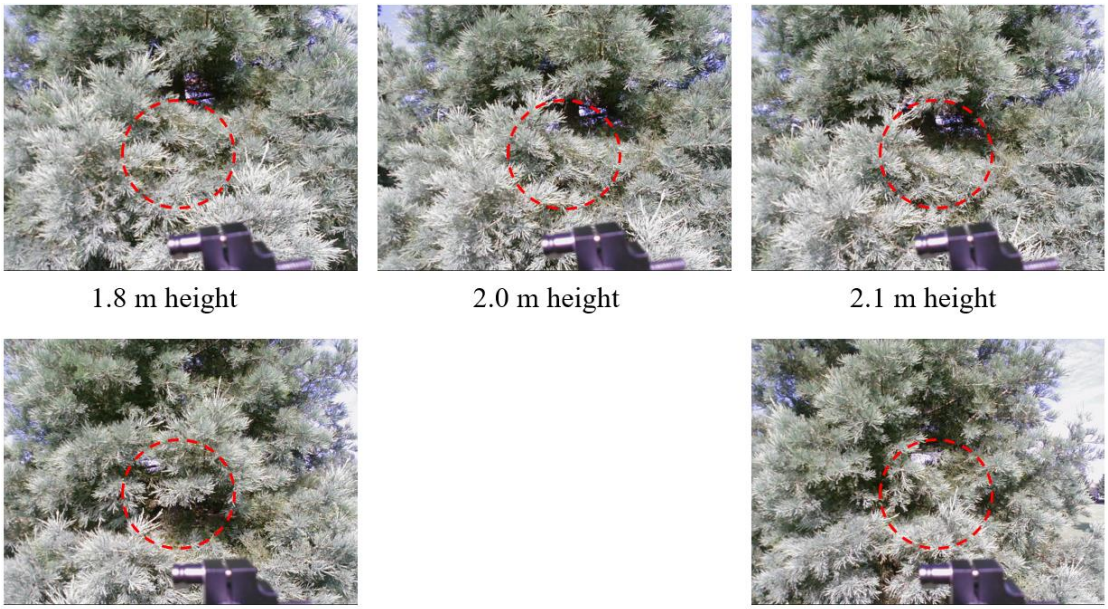
4.2.2 Impact of concave and convex shapes in crown foliage

It was found that concave crown contours decreased the τR levels, while the convex contours maximised the τR levels in various tree species. Take the in-situ test of a 5.5 m *Sequoiadendron giganteum* tree as an example. Figure 6(a) shows the τR measurement of the tree at different heights in the SAz direction (α = 37°) with or without concave shapes in the measuring patches. Viewing vision of the fiber-optic tip in tests (red circles in dashed lines represent the vision) was shown in Figure 6(b). Maximum τR appeared at 1.8 m height nearly without concave shapes in the measuring patch, while the minimum τR appeared at 2.5 m height with concave shapes accounting for approximately half of the area of the viewing vision, and the τR at other heights with concave shapes of different extents fell in between. Foliage gaps

were rather small compared to the area of concave shapes in these cases. As sunlight was captured by concave shapes in the measuring patches, the concave shapes degraded τR levels to different extents depending on the specific scenarios. When the center of the viewing vision (the circles in Figure 6(b)) deviated from the concave shapes and the ratio of concave shapes was not big, it only slightly degraded the τR (see τR spectra at heights of 2.0 m and 2.1 m in Figure 6(a)). It implies that received light of the fiber-optic tip is not evenly contributed by the viewing vision.



(a)



(b)

Figure 6. Tree crown τR measurement of a *Sequoiadendron giganteum* at different heights with or without concave shapes in the measuring patches (a) τR spectra (α = 37°); (b) viewing vision of the fiber-optic tip in tests (see red circles in dashed lines).

The impact of sparse foliage with gaps and concave crown shapes on the transflectance is complicated and heterogeneous, and different trees within a species do not share a common feature of local foliage distributions. Hence, it is preferable to avoid sampling measuring patches that exhibit sparse foliage with gaps and concave contours when collecting data for statistical analysis of radiative performance levels of various tree species. Any such patches are likely to be identified as outliers in transflectance sampling. Generally, it is possible to find measuring patches on trees with relatively dense foliage and without visible concave contours. In our field tests of 10 tree species, even for the tree species with the sparsest foliage (*Betula pendula* - silver birch) it was

easy to find dense foliage for tests without gaps in foliage and concave contours using the naked-eye.

4.2.3 Impact of solar altitude on tree crown τR

To explore the relationship between change in transflectance and solar time, solar altitude (α) was recorded when sampling the transflectance, referring to a website [72]. Figure 7 shows the τR spectra in the frontal sunlit area of a *Carpinus betulus* tree at different solar altitudes. It suggests that the tree crown transflectance increases as solar altitude rises. In further data analysis, it is found that the τR in the IR region tends to be linearly correlated with α, while no obvious variation is observed in the VIS region due to a low τR level. To dig out laws of change of τR versus α, mean transflectance in the wavelength range of 800–900 nm ($\tau R_{mean,800-900}$) was taken as an indicator, because tree crown τR spectra usually tend to be flat and hold the maximum spectral transflectance in the NIR wavelength range of 800–900 nm. Figure 8 shows the $\tau R_{mean,800-900}$ of the *Carpinus betulus* trees in terms of the tree crown τR in the frontal sunlit area linearly correlated with solar altitude α. The rule of change of IR transflectance versus solar altitude allows making a benchmark for intraspecies and interspecies comparisons in section 4.4, by converting τR spectra at different solar altitudes into equivalent τR spectra at the same solar altitude. Figure 9 shows the linear fitting results of the τR spectra in the frontal sunlit area of the *Carpinus betulus* trees varying with solar altitude in the wavelength range of 350–1000 nm.

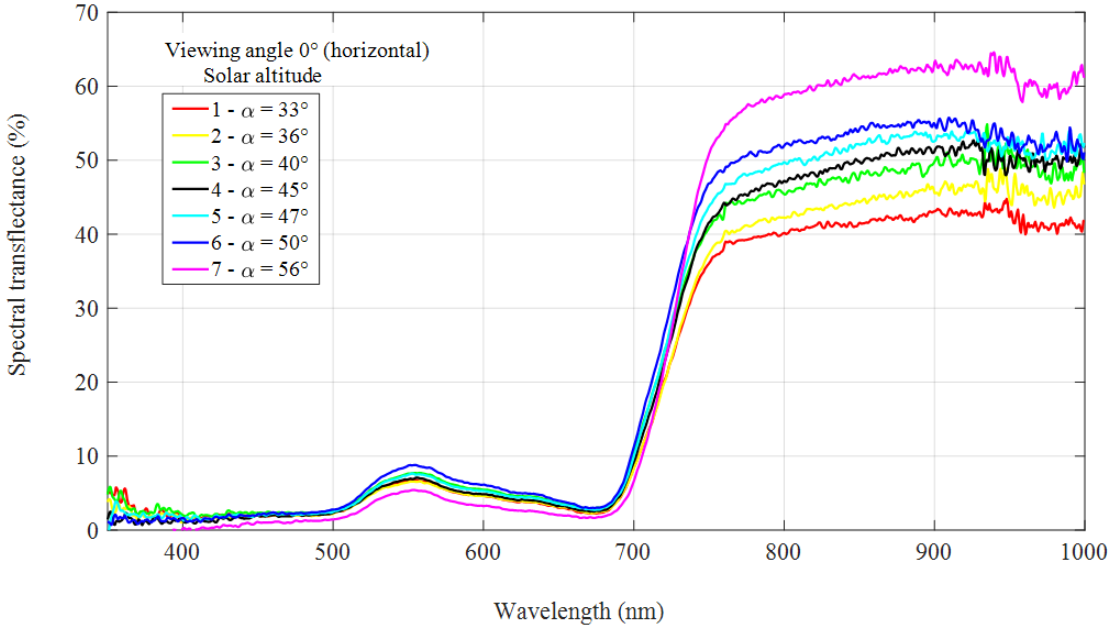


Figure 7. Transflectance spectra of tree crown surfaces for a *Carpinus betulus* tree in the frontal sunlit area at different solar altitudes (α).

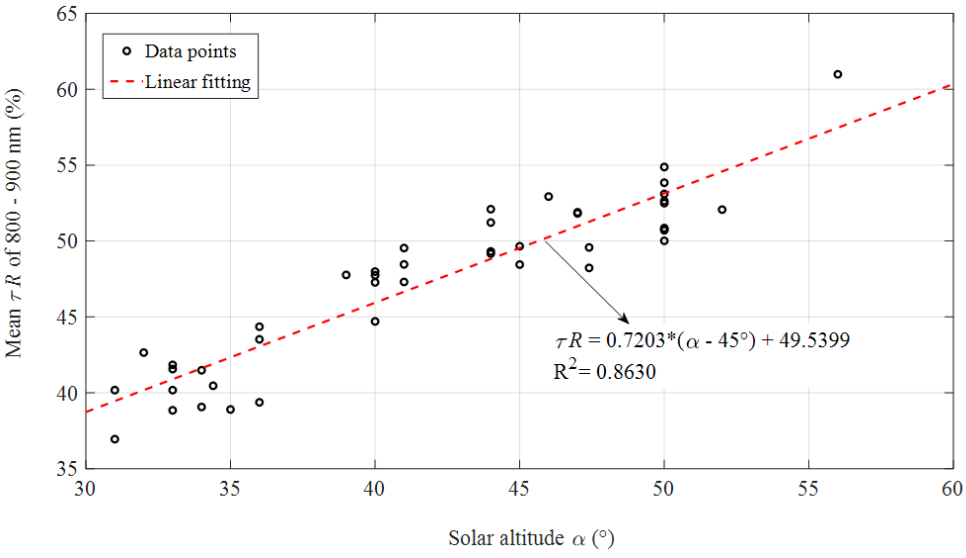
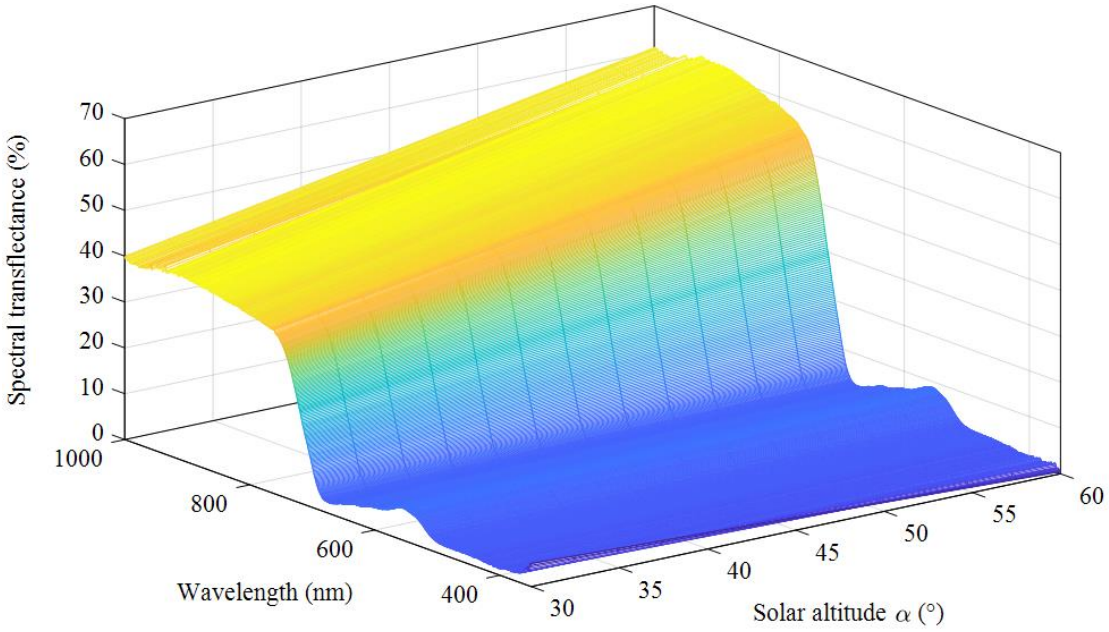
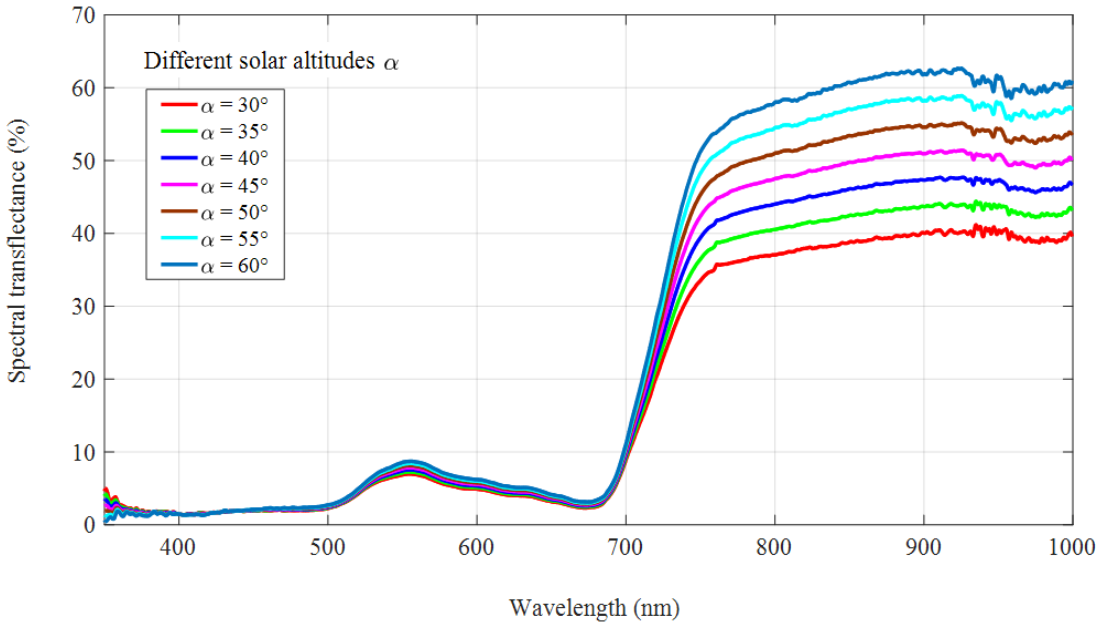


Figure 8. Linear fitting of the 800–900 nm mean transflectance ($\tau R_{mean,800-900}$) with the solar altitude α for individual *Carpinus betulus* trees using samples in the frontal sunlit area of trees.



(a) 3D plot



(b) 2D plot with discrete values of solar altitude

Figure 9. Linear fitting of tree crown transflectance spectra in the frontal sunlit area of *Carpinus betulus* trees varying with solar altitude in the wavelength range of 350–1000 nm.

In all the ten species measured, it was found that the tree crown IR transflectance in the frontal sunlit area of trees for a specific species was linearly correlated with solar altitude on sunny days. The coefficients of determination (R^2) in most cases of linear fittings were above 0.76, except the tree species *Betula pendula* and *Aesculus hippocastanum*. It was presumed that the low R^2 in the fitting of *Betula pendula* was mainly due to apparently sparse leaf density of the species. As to the species *Aesculus hippocastanum*, the low R^2 in the fitting was attributed to the development of horse chestnut leaf-miner throughout the summer [73]. Horse chestnut leaf-miner caused brown blotch mines to develop between the leaf veins, resulting in a degradation of the transflectance spectrum especially in 750–900 nm wavelength range.

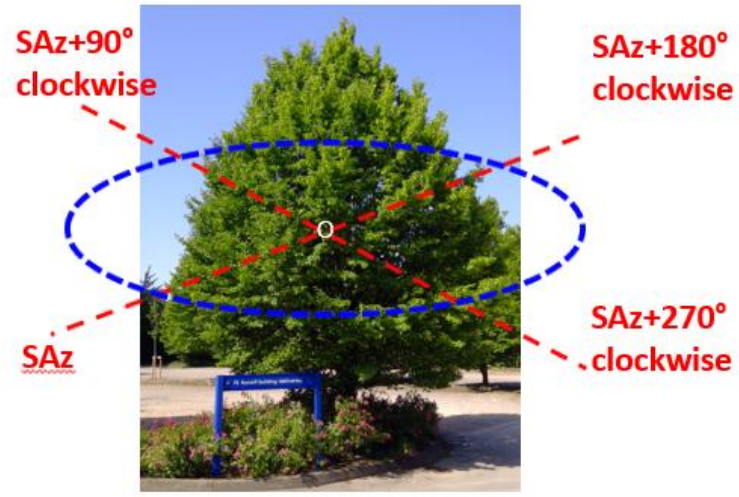
4.3 Spatial distribution of transflectance around tree crowns

As argued in our earlier work [62], transmission and reflection spectra at the leaf level were similar between trees (similar results also shown in section 4.1), while substantial variations were found in tree crown τR (transflectance) spectra due to crown structural difference and solar time. To explore spatial distribution rules of the radiative performance across a tree crown, τR spectra in typically horizontal and vertical loops around the tree crowns have been examined.

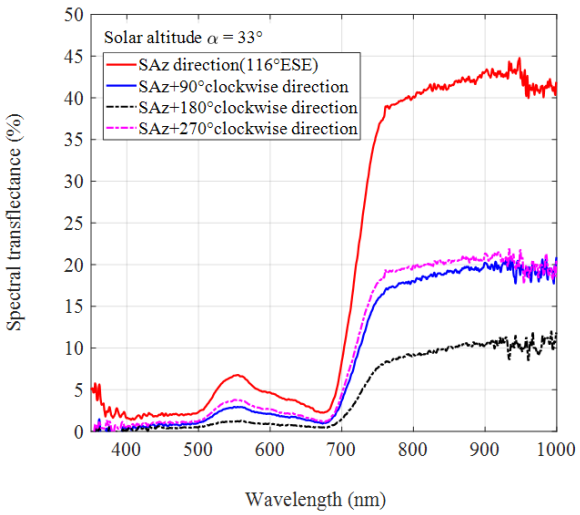
4.3.1 τR Distribution in a horizontal loop

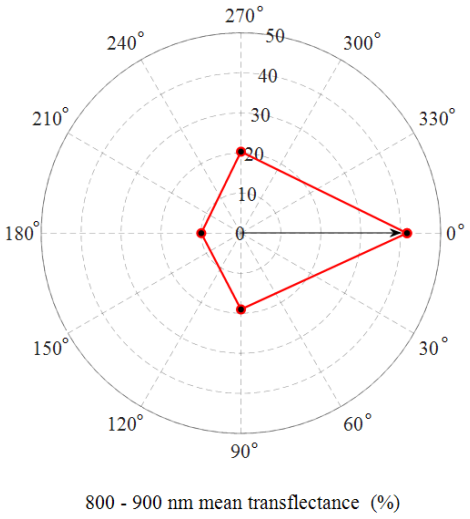
Four orientations in a horizontal plane were concerned, as shown in Figure 10(a) with the 7.0 m height *Carpinus betulus* tree. A single vertical reference plane in the solar

azimuth (SAz) direction was chosen for τR spectra measurements in the four horizontal directions (SAz; SAz +90° clockwise; SAz +180° clockwise; SAz +270° clockwise). The test sequence was implemented at 4.4 m height during 9:44–9:47 am (British Summer Time) on the sunny day of 20th August 2019 with solar altitude α = 33°. Figure 10(b) shows measured τR spectra in the four horizontal directions. To display the transflectance distribution in different directions intuitively, the $\tau R_{mean,800-900}$ (800–900 nm mean transflectance) was taken as an indicator again. Figure 10(c) plots the distribution of $\tau R_{mean,800-900}$ in the horizontal loop in polar coordinates, where the radius from the origin '0' to the point represents the value of $\tau R_{mean,800-900}$. The maximum τR spectrum in the horizontal loop appears in the frontal sunlit area of the tree, followed by SAz+90° and SAz+270° clockwise directions. The tree crown τR on the shade side (SAz +180° clockwise) has the lowest τR level. It seems that the $\tau R_{mean,800-900}$ in the SAz+90° and SAz+270° directions tends to be spatially symmetric. This was because no obvious concave contours were viewed on the measuring patches in both directions. We have observed scenarios of non-symmetric distribution in the two directions in an *Acer campestre* tree during 10:35–10:40 am on 27th June 2019, as shown in Figure 11. The $\tau R_{mean,800-900}$ in the SAz+270° clockwise direction was much lower than that in the SAz+90° clockwise direction, as apparent concave contours appeared in the SAz+270° clockwise direction, resulting in a reduction of the τR level.



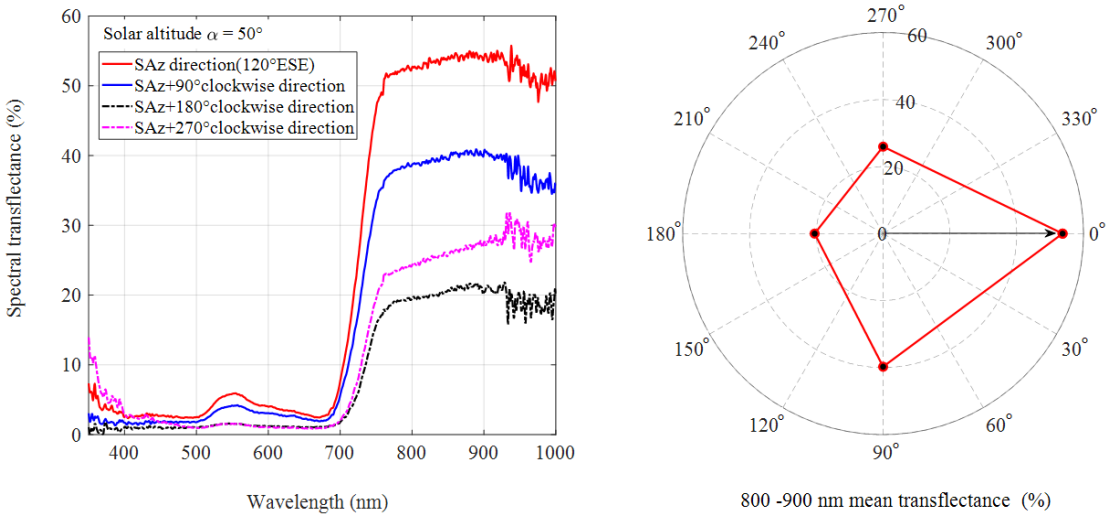
(a)





(b) (c)

Figure 10. Transflectance (τR) spectra distribution of a *Carpinus betulus* tree in a horizontal loop around the tree crown at α = 33° (a) sketch of four horizontal directions; (b) measured τR spectra in four directions (SAz – Solar azimuth; ESE – East South East); (c) distribution of $\tau R_{mean,800-900}$ in polar coordinates (0° – SAz direction; 90° – SAz +90° clockwise; 180° – SAz +180° clockwise; 270° – SAz +270° clockwise).



582

583 (b) (c)

584 **Figure 11.** Transflectance (τR) distribution of an *Acer campestre* tree in a horizontal

585 loop around the tree crown at α = 50° (a) measured τR spectra in four directions; (b)

586 distribution of $\tau R_{mean,800-900}$ in polar coordinates.

587

588 **4.3.2 τR Distribution in a vertical loop around the tree crown and aligned with**

589 **the solar azimuth direction**

590 Two typical vertical loops around the tree crown were chosen in determining

591 distribution of tree crown τR (transflectance) in vertical planes. One was a vertical

592 loop aligned with the SAz (solar azimuth) direction. The other was a vertical loop

593 perpendicular to the SAz direction (see next section 4.3.3). Figure 12(a) sketches 10

594 sampled directions distributed in the vertical loop aligned with the SAz in the 7.0 m

595 height *Carpinus betulus* tree. Figure 12(b) gives measuring results of the τR spectra

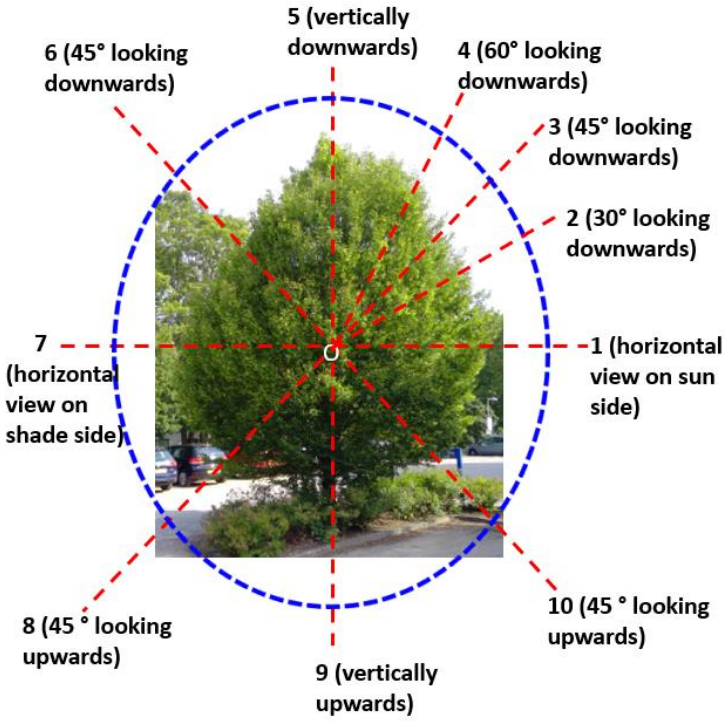
596 in the 10 sampled directions with the tree at α = 35°. The distribution of

597 $\tau R_{mean,800-900}$ is delineated in Figure 12(c). It indicates that points '1', '2', '3', '4',

598 '5' in the frontal sunlit area of the tree have high τR levels, while other points hold

relatively low τR levels. Especially, point '9' beneath the tree has the minimum $\tau R_{mean,800-900}$ value (only 6.9%). Additionally, the point '4' representing the case of fiber-optic tip tilted 60° looking downwards has the maximum $\tau R_{mean,800-900}$. The values of $\tau R_{mean,800-900}$ in the fontal sunlit area at different viewing angles (VA = 30°, 45°, 60° and 90° looking downwards) varied with solar time, as it was found that the maximum value did not maintain in the direction of VA = 60° downwards. Figure 13 gives the vertical loop distribution of the *Carpinus betulus* tree at the solar altitude α = 49°. Comparing the horizontal samples of point '1' in Figures 12(c) and 13, it confirmed that the tree crown τR at a higher solar altitude (α = 49°) is higher than that at a lower case (α = 35°). Furthermore, the strongest tree crown transflection was found primarily towards sky on the sunlit side of trees rather than towards zenith. Infrared transflection towards surrounding buildings and pedestrians is substantial. The finding provides insights on understanding radiative interactions between urban trees and surrounding built environments.

Additionally, it is noteworthy that tree crown morphology is linked to the space of the frontal sunlit area of trees, implying that tree morphology affects distributions of the tree crown τR.



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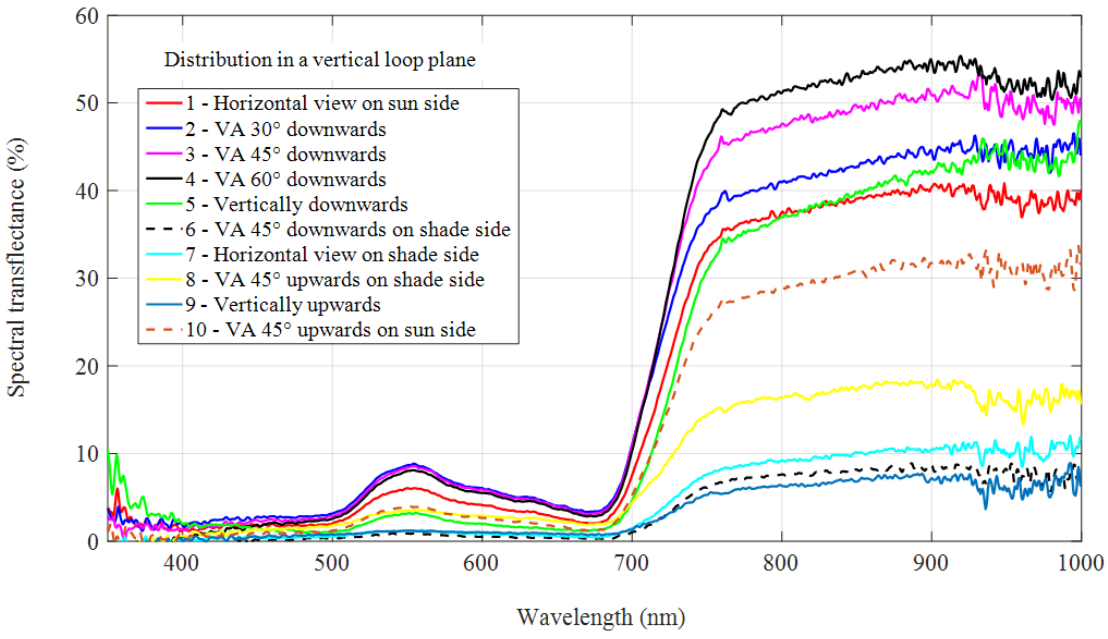
618 (a)



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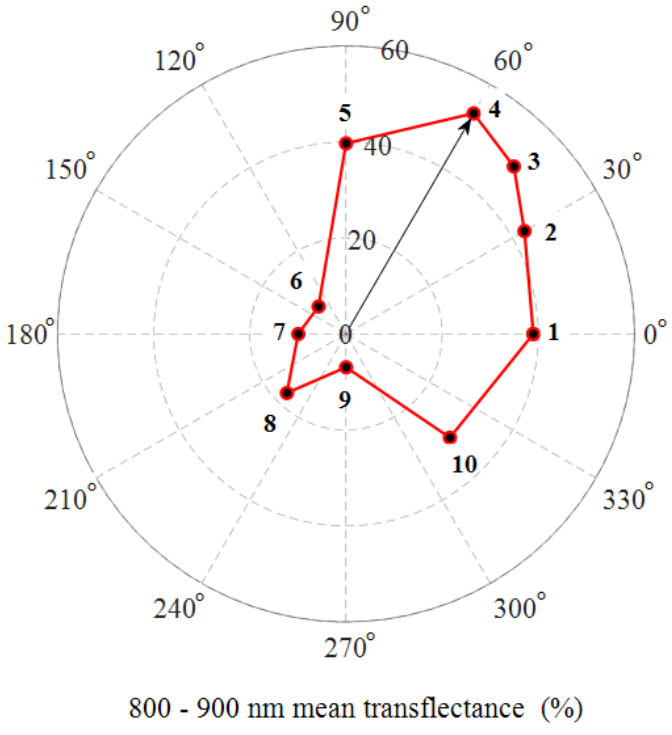
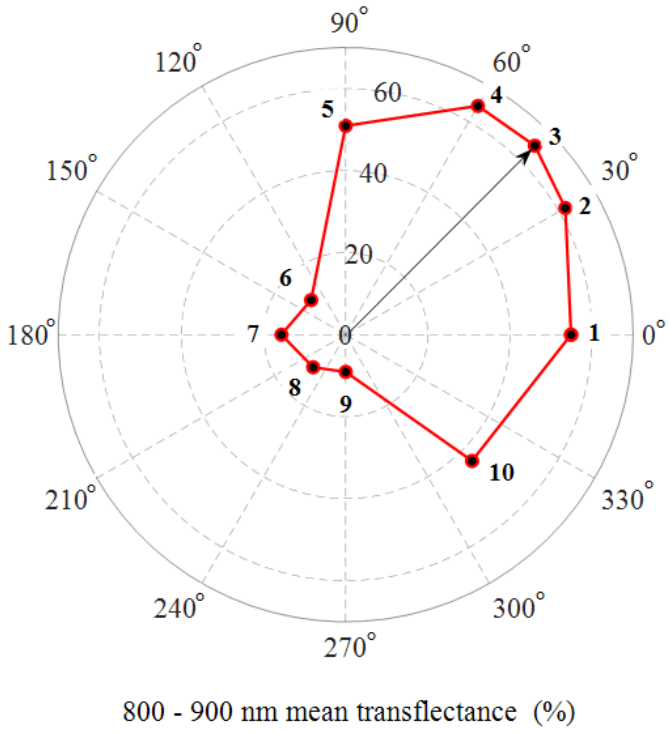
620 (b)



(c)

Figure 12. Transflectance (τR) spectra distribution of a *Carpinus betulus* tree at α = 35° in a vertical loop around tree crown and aligned with the solar azimuth (SAz) direction (a) sketch of 10 sampled directions in the vertical loop; (b) measured τR spectra; (c) distribution of $\tau R_{mean,800-900}$ in polar coordinates.



628

629 **Figure 13.** $\tau R_{mean,800-900}$ distribution of a *Carpinus betulus* tree in polar coordinates

630 at α = 49° in a vertical loop around the tree crown and aligned with the SAz direction

631

632 **4.3.3 τR Distribution in a vertical loop around the tree crown and perpendicular**
633 **to the solar azimuth direction**

634 The vertical loop perpendicular to the SAz direction was sampled in 8 directions, as

635 sketched in Figure 14(a). Figure 14(b) gives the measured τR spectra of the *Carpinus*

636 *betulus* tree in the 8 sampled directions at α = 37°, while Figure 14(b) plots

637 distribution of $\tau R_{mean,800-900}$ (800–900 nm mean) in polar coordinates. The

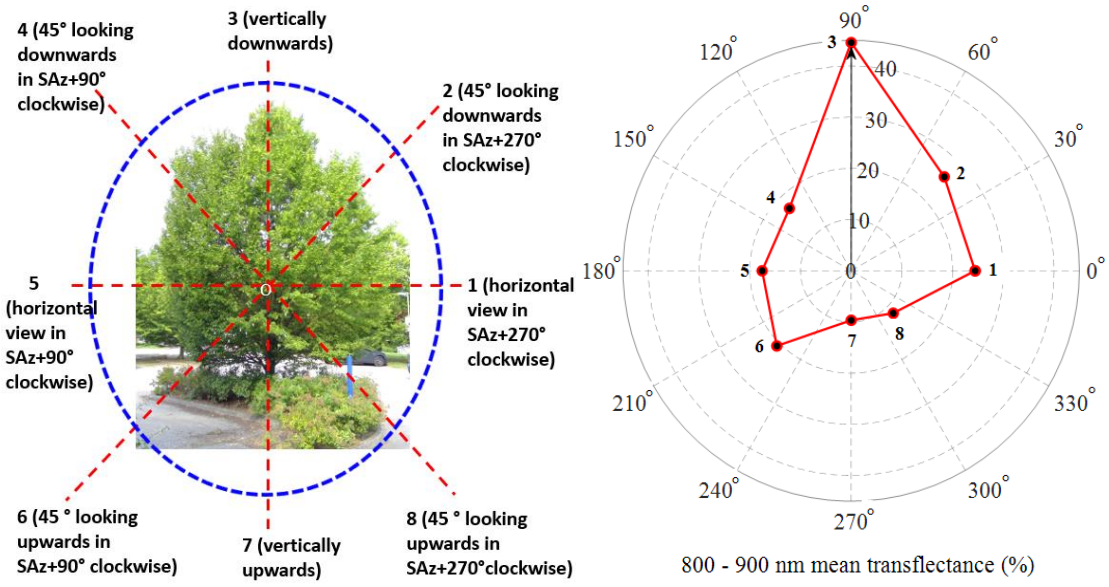
638 maximum τR appears at point '3' (vertically downwards), with τR at the other points

639 ('1', '2', '4–8') on lower levels. Comparing Figure 14(b) with the distribution in Figure

640 12(c) neglecting a slight change of solar altitude, it confirms that the primary tree crown

641 transflection is towards sky on the sunlit side of trees rather than towards zenith.

642



(a) (b)

Figure 14. Transflectance (τR) spectra distribution of a *Carpinus betulus* tree at α = 37° in a vertical loop around tree crown and perpendicular to the SAz direction (a) sketch of 8 sampled directions in the vertical loop;; (b) distribution of $\tau R_{mean,800-900}$ in polar coordinates.

4.4 Interspecific difference comparison of radiative performance levels

4.4.1 Principles of sample selections in estimating statistical mean τR spectra of tree species

As elucidated in sections 4.2 and 4.3, the impact factors of the tree crown τR are diverse, and the spatial distribution of the τR in tree crowns is nonuniform and varies with solar altitude. A benchmark is therefore needed to make interspecific comparison. Principles of test sample selections in estimating the statistical mean τR spectra of various tree species are stated below:

- Visually dense measuring patches without visible gaps in crown foliage and

concave crown contours in the viewing vision of the fiber-optic tip were selected. All trees used in the experiment were young adult to mature adult trees. Tree ages were not accounted for – the only requirement was that dense patches of foliage could be located on the tree crowns.

- The τR spectra in the frontal sunlit area of trees with vertical reference planes normal to transient solar azimuth (SAz) directions were sampled to examine intraspecific and interspecific statistical means.
- τR spectra samples with horizontal views of the fiber-optic tip were selected for statistical mean.
- Solar altitude $\alpha = 45°$ was chosen as a benchmark condition to compare interspecies difference. Measured transflectance spectra at solar altitudes different from 45° were converted to equivalent transflectance spectra at $\alpha = 45°$, considering linear correlations of transflectance spectra with α within species (referring to section 4.2.3).
- Usually, at least 3 effective transflectance spectra were sampled for an individual tree.
- At least 5 trees were sampled within a species (except the *Acer platanoides* as only four individual trees were considered) to estimate intraspecific statistical mean.
- Through chlorophyll fluorescence and heat stress measurements of part of trees selected in the field tests, it showed that the trees were subject to mild or moderate physiological stress and no significant effect of the tree physiological

stress on the τR was observed. Similarly, no significant effect of different urbanised settings of tested trees (planted on paved/sealed surfaces or on green lawns/parks) on the τR was found. Thus, the impact of heat stress and different urbanized settings of trees was not assessed in intraspecific statistical analysis.

4.4.2 Interspecific difference of infrared radiative performance levels

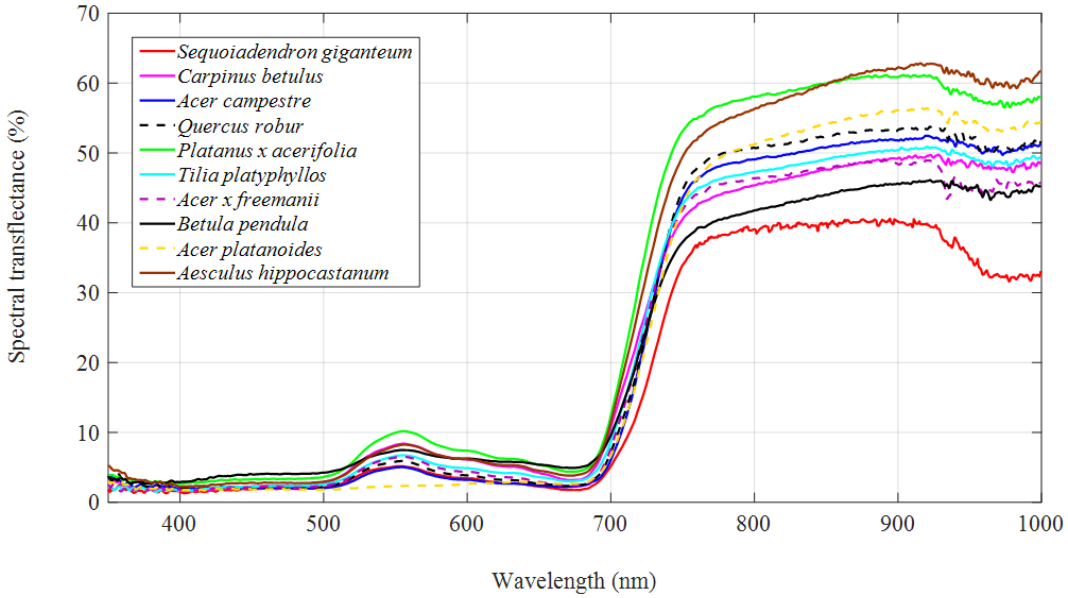
Here we use *Carpinus betulus* as an example to display the process of estimating statistical mean τR spectrum within a species. In total, effective τR spectrum samples of 9 *Carpinus betulus* trees were obtained. τR spectra of each individual tree were converted to the τR spectrum at α = 45°, and were then averaged to get a representative τR spectrum for the individual tree. The statistical mean τR spectrum of the *Carpinus betulus* species was estimated by using 9 representative τR spectra from 9 individual trees. Representative τR spectra of 5–9 trees were obtained for various tree species except the *Acer platanoides*, as only four individual trees were accessible for the copper Norway maple. Statistical mean τR spectra in the frontal sunlit area of the 10 tree species at α = 45° can be accessed in ***Appendix A***.

Figure 15 gathers the mean τR spectra and corresponding standard error bands of the 10 species, displaying interspecific performance difference. Table 1 lists statistical $\tau R_{mean,800-900}$ (800–900 nm mean transflectance) values for the 10 tree species. It suggests that *Aesculus hippocastanum* and *Platanus* x *acerifolia* have the highest levels of IR radiative performance ($\tau R_{mean,800-900}$: 0.597±0.027 for *Aesculus hippocastanum*

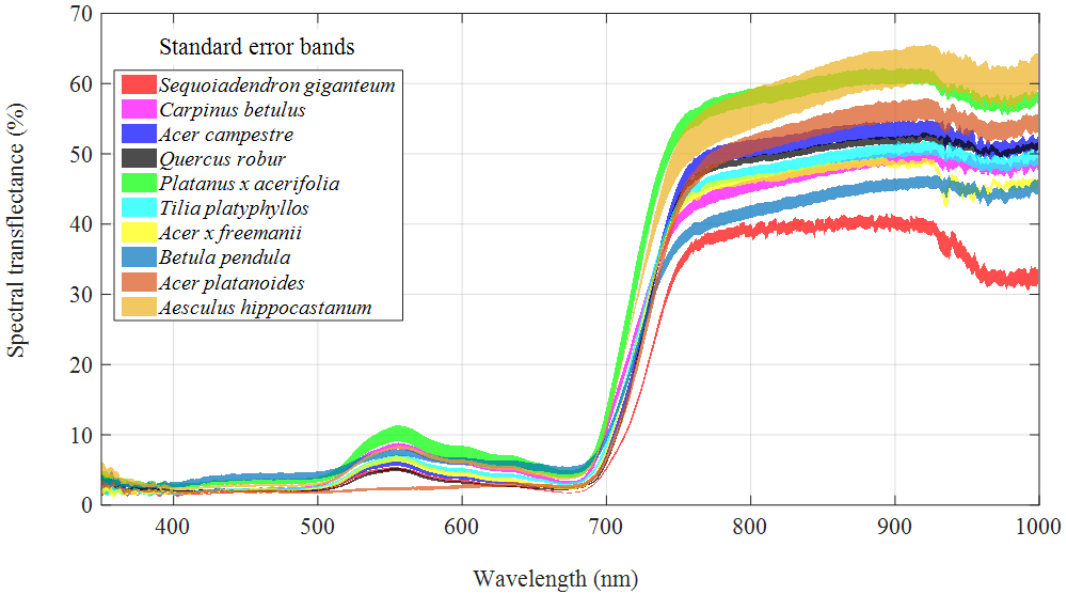
and 0.598±0.011 for *Platanus* x *acerifolia*), followed by *Acer platanoides*, *Quercus robur*, *Acer campestre*, *Tilia platyphyllos*, *Carpinus betulus* (fastigiate hornbeam) and *Acer* x *freemanii* successively on moderate levels ($\tau R_{mean,800-900}$: 0.475–0.540 and standard error within ±0.014). *Betula pendula* has the second lowest levels of IR radiative performance ($\tau R_{mean,800-900}$: 0.439±0.009), presumably due to the fact that this native British tree species is usually observed with a small leaf size and apparently sparse leaf density. *Sequoiadendron giganteum* has the minimum levels of IR radiative performance ($\tau R_{mean,800-900}$: 0.398±0.009), resulting from the minimum leaf size (needle leaves) among the tested species.

Figure 16 shows images of tree leaves for the tested species corresponding to the radiative performance levels. Interesting is that, the tree species with large-sized leaves (such as *Platanus* x *acerifolia*, *Aesculus hippocastanum*) lead to maximum radiative performance levels, followed by the tree species with moderate-sized leaves (e.g. *Acer platanoides*, *Quercus robur*, *Acer campestre*, *Tilia platyphyllos*). *Carpinus betulus* has a leaf size smaller than other tree species on the moderate levels of radiative performance, resulting in a slightly lower performance level than the others except *Acer* x *freemanii*. Although the leaf size of *Acer* x *freemanii* seems to be close to that of *Quercus robur* and *Acer campestre*, branching structure of its leaves degrades its radiative performance level. From this perspective, each individual leaf of *Acer* x *freemanii* can be viewed as three leaflets which have similar leaf sizes as the *Carpinus betulus* leaves, resulting in the same radiative performance level as the latter.

Sequoiadendron giganteum with the smallest needle leaves among the 10 species has the minimum radiative performance level. It reveals that interspecific difference of infrared radiative performance levels strongly depends on leaf size when visibly dense foliage (no visible gaps and concave shapes in crown foliage) is observed on the tree crown contours. The finding confirms the viewpoint in [74] that species average leaf size is the most important determinant of self-shading that affected light capture. It provides insights on species selection for heat stress mitigation in urban microclimates. The impact of leaf size implies that disregarding tree ages (young or old) in sampling visually dense measuring patches (see 4.4.1) is appropriate, as mature trees of a specific-species tend to have reasonably similar leaf shape and leaf size.



(a)

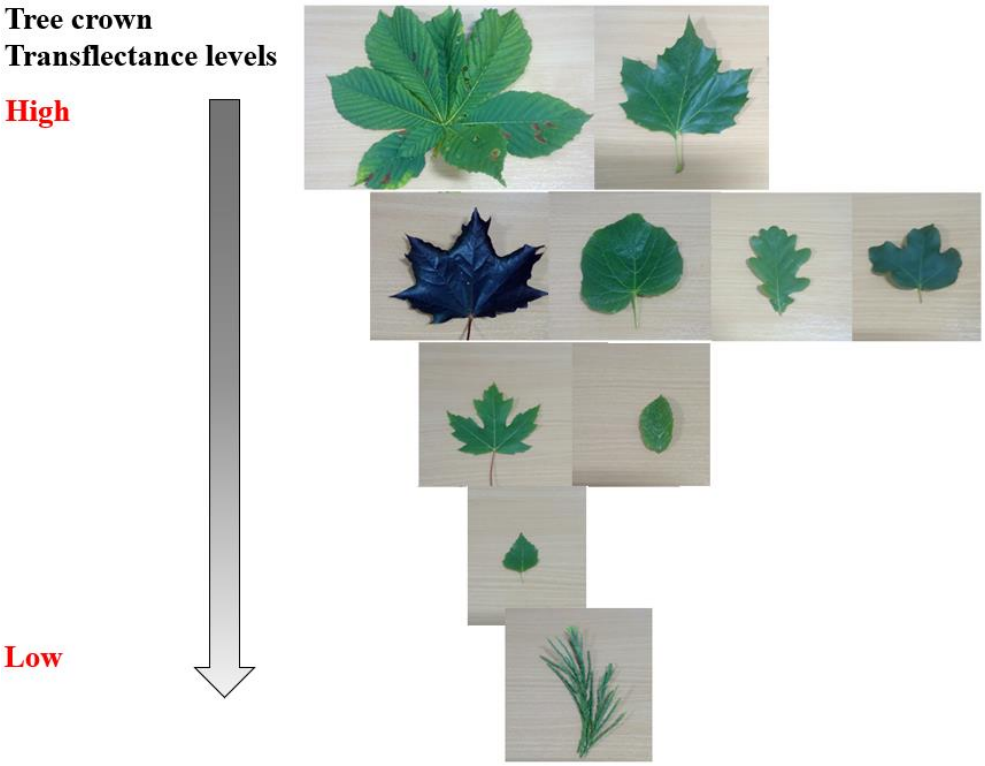


(b)

Figure 15. Interspecific radiative performance difference of ten tree species planted in the UK in terms of equivalent τR spectra at α = 45° (a) statistical mean τR spectra; (b) standard mean error bands.

Table 1. Statistical $\tau R_{mean,800-900}$ for ten tree species

Tree species	$\tau R_{mean,800-900}$ (%)	Standard mean error (%)
Sequoiadendron giganteum (Giant sequoia)	39.8	± 0.9
Carpinus betulus (Fastigiate hornbeam)	47.5	± 0.9
Acer campestre (Field maple)	52.2	± 1.2
Quercus robur (English oak)	50.7	± 0.5
Platanus x *acerifolia* (London plane)	59.8	± 1.1
Tilia platyphyllos (Large-leaved lime)	49.0	± 0.9
Acer x *freemanii* (Autumn blaze maple)	47.8	± 0.6
Betula pendula (Silver Birch)	43.9	± 0.9
Acer platanoides (Copper Norway maple)	54.0	± 1.4
Aesculus hippocastanum (Horse chestnut)	59.7	± 2.7



744

745 **Figure 16.** Radiative performance levels of multiple tree species classified by leaf size.

746

747 **4.4.3 Identification of intraspecific and interspecific differences in canopy**
748 **transflectance by PCA**

749 At the scale of crown patches significant differences in transflectance were observed
750 between species, and these corresponded to different biological and spectral properties.
751 PCA of the 5 nm spectral bands (n = 130 per tree crown) collected from 67 individual
752 trees across 10 species revealed significant interspecific differences in crown
753 transflectance for the 350-1000 nm range. The first two PCA axes explained 86.9% of
754 the variance in spectral profiles and were significantly correlated with different canopy
755 transflectance properties (see Table 2).

756

757 **Table 2.** Relationship between first five PCA axes and bands of canopy transflectance

758 in the 350-1000 nm range across 67 individual trees drawn from 10 species

Axis	Variance explained	Cumulative variance	Positive correlation (*r*)	Negative correlation (*r*)
1	54.5%	54.5%	All except 370-375 nm	
2	32.4%	86.9%	735-1000 nm	395-705 nm
3	7.2%	94.1%	350-485 nm	525-615 & 695-735 nm
4	2.6%	96.7%	350-400 nm	
5	1.5%	98.2%		350-355 nm

759

760 Increasing transflectance in the 350-1000 nm range corresponded to axis 1 of the PCA
761 and was the most explanatory factor contributing to species differences (Table 2). The
762 strongest correlation to axis 1 was found in the transition between VIS and NIR; large-
763 leaved species (*Platanus* x *acerifolia* and *Aesculus hippocastanum*) were positively
764 associated with this axis, whereas the only conifer (*Sequoiadendron giganteum*) was
765 negatively associated (Species $R^2 = 0.731$; $P < 0.001$; Figure 17). Species were further
766 sub-divided by their opposing interactions with NIR and components of VIS. Axis 2
767 corresponded with increasing NIR and decreasing VIS (violet, blue, yellow and orange)
768 transflectance (Table 2); *Acer platanoides* and *Quercus robur* were positively
769 associated with this second axis, whereas *Sequoiadendron giganteum* and *Betula*
770 *pendula* were negatively associated (Species $R^2 = 0.663$; $P < 0.001$; Figure 17). These
771 results show that interspecific differences between the transflectance profiles of
772 commonly planted tree species are readily measurable and reveal key differences
773 between species that can alter radiative performance in urban areas. As noted by
774 Cavender-Barres *et al.* [70], leaf spectra can be considered as an integrated measure of
775 phenotype; hence, further research into the impacts of environmental stress on canopy
776 transflectance could help in understanding the likely consequences of climate change

for urban areas, in terms of the impact on radiative performance and energy balance.

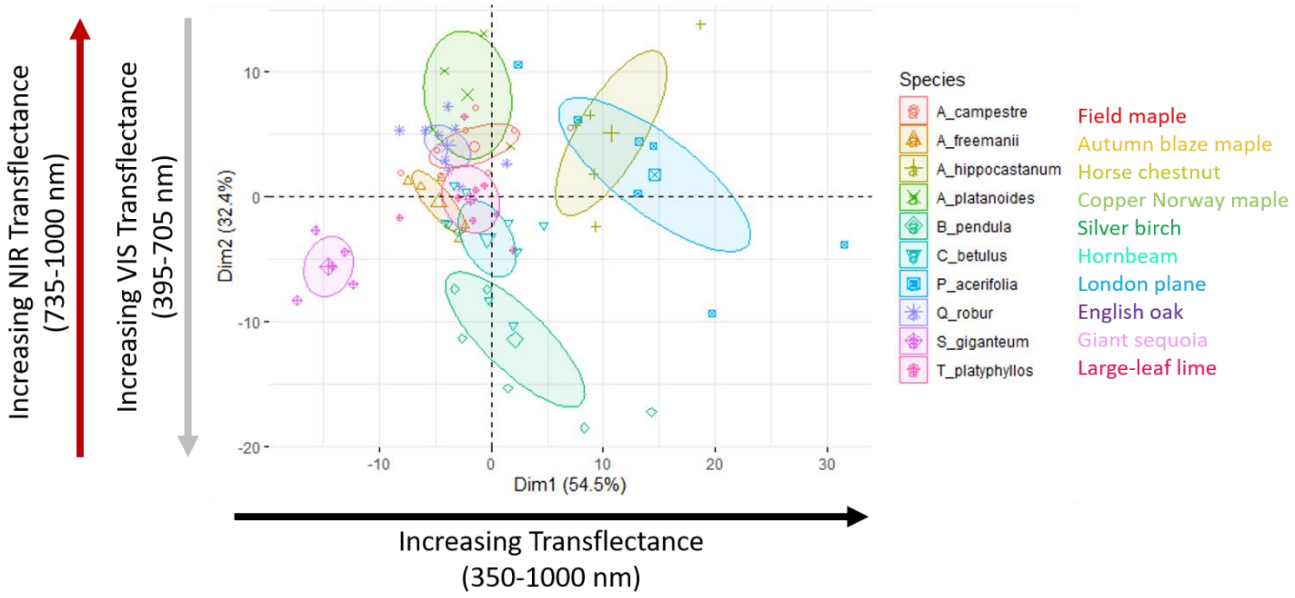


Figure 17. PCA of crown transflectance spectra for axes 1 and 2 displaying differentiation between species and spectral properties correlated with each axis. Polygons represent 95% confidence interval for each species.

Furthermore, PERMANOVA using distance indices supported significant interspecific variation in crown transflectance (Species, 999 permutations: *Pseudo-F* $_{df\ 9,57}$ = 14.3; R^2 = 0.694; *P* = 0.001). Although variability in crown transflectance was observed within species, examination of the homogeneity of multivariate dispersion indicated that intraspecific variance (i.e. variance within species) was not significantly different between species (Species, 999 permutations: $F_{df9,57}$ = 1.86; *P* = 0.077). Hence, crown transflectance values were relatively similar within species, and the degree of multivariate variability within species was relatively similar between species. However, this observation was based on collecting spectral data from canopy patches with a

consistent, generally convex, coverage of leaves and avoiding gaps or concave patches. Therefore, intraspecific differences are expected to be strongly influenced by their unique life-history, including interactions with the biotic/abiotic environment and local government landscaping and management decisions.

5 Conclusions

Based on a tree crown spectroscopy measurement method established earlier, substantial in-situ tests of radiative performance of 10 tree species have been implemented in terms of the tree crown transflectance (τR). Spatial distribution rules of the τR across tree crowns were identified. Infrared radiative performance difference of the 10 tree species (*Sequoiadendron giganteum*, *Carpinus betulus*, *Acer campestre*, *Quercus robur*, *Platanus* x *acerifolia*, *Tilia platyphyllos*, *Acer* x *freemanii*, *Betula pendula*, *Acer platanoides*, *Aesculus hippocastanum*) commonly planted in the UK was statistically determined in terms of τR spectra in frontal sunlit area of the trees, by converting τR spectra on the same benchmark of solar altitude α = 45°. Main findings are as follows:

- Mean leaf reflectance spectra of various tree species have a minor difference between each other, with spectral reflectance deviations of ± 4%, contrasting to substantial differences in spatial distribution of the τR across the tree crowns.
- Impact factors of tree crown transflectance (sparse foliage with gaps in crown foliage, concave and convex shapes in crown foliage, solar time): Visibly non-uniform foliage distribution in the measuring patches, such as sparse foliage

with gaps and concave shapes degrades τR levels to different extents. Impact of sparse foliage with gaps and concave shapes on τR is complicated and heterogeneous. τR in the frontal sunlit area of trees for a specific species is linearly correlated with solar altitude on sunny days, allowing this to be used as a benchmark for comparing differences in intraspecific and interspecific performance.

- Spatial distribution rules: The primary tree crown τR in a horizontal loop around tree crowns appears in the frontal sunlit area (SAz direction), followed by those in SAz+90° and SAz +270° clockwise directions. The τR on the tree shade side (SAz+180° clockwise) has the lowest levels). τR distributions in two typically vertical loops perpendicular to each other confirm that the strongest tree crown transflection was found primarily towards sky on the sunlit side of trees rather than towards zenith. The direction of the maximum transflection in the frontal sunlit area varies with solar time and depends on the polar angle consisting of solar azimuth and solar altitude. The τR beneath trees has the minimum level.
- Interspecific difference comparison of infrared radiative performance levels indicates that tree species such as *Platanus* x *acerifolia* and *Aesculus hippocastanum* with large-sized leaves, lead to maximum radiative performance levels, followed by the tree species with moderate-sized leaves (e.g. *Acer platanoides*, *Acer campestre*, *Quercus robur*, *Tilia platyphyllos*). *Sequoiadendron giganteum* has the minimum radiative performance level,

mainly due to its small needle leaves. It reveals that interspecific difference of the infrared radiative performance levels strongly depends on leaf size if visually dense foliage (no obvious gaps in foliage and no concave shapes) is observed on the tree crown contours.

The findings provide insights on understanding radiative interactions between urban trees and the surrounding built environments.

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Appendix A. Supplementary materials

Supplementary data associated with this article is openly available in the University of Reading Research Data Archive, in the online version, at: http://dx.doi.org/10.17864/1947.231. Metadata will be available by request.

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References

[1] R. Garcia-Herrera, J. Díaz, R.M. Trigo, J. Luterbacher, E.M. Fischer, A review of the european summer heat wave of 2003, Crit. Rev. Environ. Sci. Technol. 40 (2010) 267–306. https://doi.org/10.1080/10643380802238137.

[2] IPCC, Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, 2014.

[3] A. Gasparrini, B. Armstrong, The impact of heat waves on mortality, Epidemiology. (2011). https://doi.org/10.1097/EDE.0b013e3181fdcd99.

[4] Y. Guo, A. Gasparrini, B.G. Armstrong, *et al.*, Heat wave and mortality: A multicountry, multicommunity study, Environ. Health Perspect. (2017). https://doi.org/10.1289/EHP1026.

[5] E.M. Fischer, R. Knutti, Anthropogenic contribution to global occurrence of heavy-precipitation and high-temperature extremes, Nat. Clim. Chang. 5 (2015) 560–564. https://doi.org/10.1038/nclimate2617.

[6] L. Zhao, M. Oppenheimer, Q. Zhu, J.W. Baldwin, K.L. Ebi, E. Bou-Zeid, K.

879 Guan, X. Liu, Interactions between urban heat islands and heat waves, Environ.
880 Res. Lett. 13 (2018) 034003. https://doi.org/10.1088/1748-9326/aa9f73.
881 [7] C.C. Konijnendijk, K. Nilsson, T.B. Randrup, J. Schipperijn, Urban forests and
882 trees: A reference book, 2005. https://doi.org/10.1007/3-540-27684-X.
883 [8] R. Upreti, Z.H. Wang, J. Yang, Radiative shading effect of urban trees on cooling
884 the regional built environment, Urban For. Urban Green. 26 (2017) 18–24.
885 https://doi.org/10.1016/j.ufug.2017.05.008.
886 [9] Z. Zou, Y. Yang, G.Y. Qiu, Quantifying the evapotranspiration rate and its
887 cooling effects of urban hedges based on three-temperature model and infrared
888 remote sensing, Remote Sens. 11 (2019) 1–18.
889 https://doi.org/10.3390/rs11020202.
890 [10] M. Taleghani, Outdoor thermal comfort by different heat mitigation strategies-
891 A review, Renew. Sustain. Energy Rev. 81 (2018) 2011–2018.
892 https://doi.org/10.1016/j.rser.2017.06.010.
893 [11] L. Kong, K.K.L. Lau, C. Yuan, Y. Chen, Y. Xu, C. Ren, E. Ng, Regulation of
894 outdoor thermal comfort by trees in Hong Kong, Sustain. Cities Soc. 31 (2017)
895 12–25. https://doi.org/10.1016/j.scs.2017.01.018.
896 [12] Z.H. Wang, X. Zhao, J. Yang, J. Song, Cooling and energy saving potentials of
897 shade trees and urban lawns in a desert city, Appl. Energy. 16 (2016) 437–444.
898 https://doi.org/10.1016/j.apenergy.2015.10.047.
899 [13] S. Gillner, J. Vogt, A. Tharang, S. Dettmann, A. Roloff, Role of street trees in
900 mitigating effects of heat and drought at highly sealed urban sites, Landsc. Urban

Plan. 143 (2015) 33–42. https://doi.org/10.1016/j.landurbplan.2015.06.005.

[14] H. Lee, H. Mayer, L. Chen, Contribution of trees and grasslands to the mitigation of human heat stress in a residential district of Freiburg, Southwest Germany, Landsc. Urban Plan. 148 (2016) 37–50. https://doi.org/10.1016/j.landurbplan.2015.12.004.

[15] A.S. Yang, Y.H. Juan, C.Y. Wen, C.J. Chang, Numerical simulation of cooling effect of vegetation enhancement in a subtropical urban park, Appl. Energy. 192 (2017) 178–200. https://doi.org/10.1016/j.apenergy.2017.01.079.

[16] D.E. Bowler, L. Buyung-Ali, T.M. Knight, A.S. Pullin, Urban greening to cool towns and cities: A systematic review of the empirical evidence, Landsc. Urban Plan. 97 (2010) 147–155. https://doi.org/10.1016/j.landurbplan.2010.05.006.

[17] W. Zhou, J. Wang, M.L. Cadenasso, Effects of the spatial configuration of trees on urban heat mitigation: A comparative study, Remote Sens. Environ. 195 (2017) 1–12. https://doi.org/10.1016/j.rse.2017.03.043.

[18] S. Sodoudi, H. Zhang, X. Chi, F. Müller, H. Li, The influence of spatial configuration of green areas on microclimate and thermal comfort, Urban For. Urban Green. 34 (2018) 85–96. https://doi.org/10.1016/j.ufug.2018.06.002.

[19] Z. Tan, K.K.L. Lau, E. Ng, Urban tree design approaches for mitigating daytime urban heat island effects in a high-density urban environment, Energy Build. 114 (2016) 265–274. https://doi.org/10.1016/j.enbuild.2015.06.031.

[20] T. Zölch, M.A. Rahman, E. Pfleiderer, G. Wagner, S. Pauleit, Designing public squares with green infrastructure to optimize human thermal comfort, Build.

923 Environ. 149 (2019) 640–654. https://doi.org/10.1016/j.buildenv.2018.12.051.
924 [21] E. Jamei, P. Rajagopalan, M. Seyedmahmoudian, Y. Jamei, Review on the impact
925 of urban geometry and pedestrian level greening on outdoor thermal comfort,
926 Renew. Sustain. Energy Rev. 54 (2016) 1002–1017.
927 https://doi.org/10.1016/j.rser.2015.10.104.
928 [22] Y. Wang, U. Berardi, H. Akbari, Comparing the effects of urban heat island
929 mitigation strategies for Toronto, Canada, Energy Build. 114 (2016) 2–19.
930 https://doi.org/10.1016/j.enbuild.2015.06.046.
931 [23] Y. Liu, D.J. Harris, Effects of shelterbelt trees on reducing heating-energy
932 consumption of office buildings in Scotland, Appl. Energy. 85 (2008) 115–127.
933 https://doi.org/10.1016/j.apenergy.2007.06.008.
934 [24] J.L. Moss, K.J. Doick, S. Smith, M. Shahrestani, Influence of evaporative
935 cooling by urban forests on cooling demand in cities, Urban For. Urban Green.
936 37 (2019) 65–73. https://doi.org/10.1016/j.ufug.2018.07.023.
937 [25] C.M. Hsieh, J.J. Li, L. Zhang, B. Schwegler, Effects of tree shading and
938 transpiration on building cooling energy use, Energy Build. 159 (2018) 382–397.
939 https://doi.org/10.1016/j.enbuild.2017.10.045.
940 [26] N. Bréda, V. Badeau, Forest tree responses to extreme drought and some biotic
941 events: Towards a selection according to hazard tolerance?, Comptes Rendus -
942 Geosci. 340 (2008) 651–662. https://doi.org/10.1016/j.crte.2008.08.003.
943 [27] C. Calfapietra, J. Peñuelas, Ü. Niinemets, Urban plant physiology: Adaptation-
944 mitigation strategies under permanent stress, Trends Plant Sci. 20 (2015) 72–75.

945 https://doi.org/10.1016/j.tplants.2014.11.001.
946 [28] R. Teskey, T. Wertin, I. Bauweraerts, M. Ameye, M.A. McGuire, K. Steppe,
947 Responses of tree species to heat waves and extreme heat events, Plant Cell
948 Environ. 38 (2015) 1699–1712. https://doi.org/10.1111/pce.12417.
949 [29] I. Bauweraerts, M. Ameye, T.M. Wertin, M.A. McGuire, R.O. Teskey, K. Steppe,
950 Water availability is the decisive factor for the growth of two tree species in the
951 occurrence of consecutive heat waves, Agric. For. Meteorol. 189–190 (2014) 19–
952 29. https://doi.org/10.1016/j.agrformet.2014.01.001.
953 [30] S. Leuzinger, R. Vogt, C. Körner, Tree surface temperature in an urban
954 environment, Agric. For. Meteorol. 150 (2010) 56–62.
955 https://doi.org/10.1016/j.agrformet.2009.08.006.
956 [31] S. Zheng, J.M. Guldmann, Z. Liu, L. Zhao, Influence of trees on the outdoor
957 thermal environment in subtropical areas: An experimental study in Guangzhou,
958 China, Sustain. Cities Soc. 42 (2018) 482–497.
959 https://doi.org/10.1016/j.scs.2018.07.025.
960 [32] M.A. Irmak, S. Yilmaz, E. Mutlu, H. Yilmaz, Assessment of the effects of
961 different tree species on urban microclimate, Environ. Sci. Pollut. Res. 25 (2018)
962 15802–15822. https://doi.org/10.1007/s11356-018-1697-8.
963 [33] C.Y. Park, D.K. Lee, E.S. Krayenhoff, H.K. Heo, S. Ahn, T. Asawa, A.
964 Murakami, H.G. Kim, A multilayer mean radiant temperature model for
965 pedestrians in a street canyon with trees, Build. Environ. 141 (2018) 298–309.
966 https://doi.org/10.1016/j.buildenv.2018.05.058.

[34] C.Y. Park, D.K. Lee, E.S. Krayenhoff, H.K. Heo, J.H. Hyun, K. Oh, T.Y. Park, Variations in pedestrian mean radiant temperature based on the spacing and size of street trees, Sustain. Cities Soc. 48 (2019) 1–9. https://doi.org/10.1016/j.scs.2019.101521.

[35] T. Zölch, J. Maderspacher, C. Wamsler, S. Pauleit, Using green infrastructure for urban climate-proofing: An evaluation of heat mitigation measures at the micro-scale, Urban For. Urban Green. 20 (2016) 305–316. https://doi.org/10.1016/j.ufug.2016.09.011.

[36] T.E. Morakinyo, K.K.L. Lau, C. Ren, E. Ng, Performance of Hong Kong's common trees species for outdoor temperature regulation, thermal comfort and energy saving, Build. Environ. 137 (2018) 157–170. https://doi.org/10.1016/j.buildenv.2018.04.012.

[37] D. Armson, M.A. Rahman, A.R. Ennos, A comparison of the shading effectiveness of five different street tree species in Manchester, UK, Arboric. Urban For. 39 (2013) 157–164.

[38] M.A. Rahman, D. Armson, A.R. Ennos, A comparison of the growth and cooling effectiveness of five commonly planted urban tree species, Urban Ecosyst. 18 (2015) 371–389. https://doi.org/10.1007/s11252-014-0407-7.

[39] L. Zhang, Q. Zhan, Y. Lan, Effects of the tree distribution and species on outdoor environment conditions in a hot summer and cold winter zone: A case study in Wuhan residential quarters, Build. Environ. 130 (2018) 27–39. https://doi.org/10.1016/j.buildenv.2017.12.014.

[40] R. Sanusi, D. Johnstone, P. May, S.J. Livesley, Microclimate benefits that different street tree species provide to sidewalk pedestrians relate to differences in Plant Area Index, Landsc. Urban Plan. 157 (2017) 502–511. https://doi.org/10.1016/j.landurbplan.2016.08.010.

[41] X. Chen, P. Zhao, Y. Hu, *et al.*, Canopy transpiration and its cooling effect of three urban tree species in a subtropical city- Guangzhou, China, Urban For. Urban Green. 43 (2019) 126368. https://doi.org/10.1016/j.ufug.2019.126368.

[42] M. Jiao, W. Zhou, Z. Zheng, *et al.*, Patch size of trees affects its cooling effectiveness: A perspective from shading and transpiration processes, Agric. For. Meteorol. 247 (2017) 293–299. https://doi.org/10.1016/j.agrformet.2017.08.013.

[43] E. Litvak, H.R. McCarthy, D.E. Pataki, A method for estimating transpiration of irrigated urban trees in California, Landsc. Urban Plan. 158 (2017) 48–61. https://doi.org/10.1016/j.landurbplan.2016.09.021.

[44] M.A. Rahman, A. Moser, A. Gold, T. Rötzer, S. Pauleit, Vertical air temperature gradients under the shade of two contrasting urban tree species during different types of summer days, Sci. Total Environ. 633 (2018) 100–111. https://doi.org/10.1016/j.scitotenv.2018.03.168.

[45] J. Konarska, J. Uddling, B. Holmer, M. Lutz, F. Lindberg, H. Pleijel, S. Thorsson, Transpiration of urban trees and its cooling effect in a high latitude city, Int. J. Biometeorol. 60 (2016) 159–172. https://doi.org/10.1007/s00484-015-1014-x.

[46] C. Wang, Z.H. Wang, C. Wang, S.W. Myint, Environmental cooling provided by urban trees under extreme heat and cold waves in U.S. cities, Remote Sens.

Environ. 227 (2019) 28–43. https://doi.org/10.1016/j.rse.2019.03.024.

[47] M.A. Rahman, A. Moser, T. Rötzer, S. Pauleit, Comparing the transpirational and shading effects of two contrasting urban tree species, Urban Ecosyst. 22 (2019) 683–697. https://doi.org/10.1007/s11252-019-00853-x.

[48] W.T.L. Chow, A.J. Brazel, Assessing xeriscaping as a sustainable heat island mitigation approach for a desert city, Build. Environ. 47 (2012) 170–181. https://doi.org/10.1016/j.buildenv.2011.07.027.

[49] A. Middel, N. Chhetri, R. Quay, Urban forestry and cool roofs: Assessment of heat mitigation strategies in Phoenix residential neighborhoods, Urban For. Urban Green. 14 (2015) 178–186. https://doi.org/10.1016/j.ufug.2014.09.010.

[50] C.Y. Park, D.K. Lee, E.S. Krayenhoff, *et al.*, Variations in pedestrian mean radiant temperature based on the spacing and size of street trees, Sustain. Cities Soc. (2019). https://doi.org/10.1016/j.scs.2019.101521.

[51] L.V. de Abreu-Harbich, L.C. Labaki, A. Matzarakis, Effect of tree planting design and tree species on human thermal comfort in the tropics, Landsc. Urban Plan. 138 (2015) 99–109. https://doi.org/10.1016/j.landurbplan.2015.02.008.

[52] Z. Tan, K.K.L. Lau, E. Ng, Planning strategies for roadside tree planting and outdoor comfort enhancement in subtropical high-density urban areas, Build. Environ. 120 (2017) 93–109. https://doi.org/10.1016/j.buildenv.2017.05.017.

[53] Z. Wu, L. Chen, Optimizing the spatial arrangement of trees in residential neighborhoods for better cooling effects: Integrating modeling with in-situ measurements, Landsc. Urban Plan. 167 (2017) 463–472.

https://doi.org/10.1016/j.landurbplan.2017.07.015.

[54] Z. Wu, P. Dou, L. Chen, Comparative and combinative cooling effects of different spatial arrangements of buildings and trees on microclimate, Sustain. Cities Soc. (2019). https://doi.org/10.1016/j.scs.2019.101711.

[55] Q. Zhao, D.J. Sailor, E.A. Wentz, Impact of tree locations and arrangements on outdoor microclimates and human thermal comfort in an urban residential environment, Urban For. Urban Green. 32 (2018) 81–91. https://doi.org/10.1016/j.ufug.2018.03.022.

[56] M. Aminipouri, D. Rayner, F. Lindberg, *et al.*, Urban tree planting to maintain outdoor thermal comfort under climate change: The case of Vancouver's local climate zones, Build. Environ. 158 (2019) 226–236. https://doi.org/10.1016/j.buildenv.2019.05.022.

[57] Z.H. Wang, Monte Carlo simulations of radiative heat exchange in a street canyon with trees, Sol. Energy. 110 (2014) 704–713. https://doi.org/10.1016/j.solener.2014.10.012.

[58] T.E. Morakinyo, Y.F. Lam, Simulation study on the impact of tree-configuration, planting pattern and wind condition on street-canyon's micro-climate and thermal comfort, Build. Environ. 103 (2016) 262–275. https://doi.org/10.1016/j.buildenv.2016.04.025.

[59] T. Eckmann, A. Morach, M. Hamilton, *et al.*, A. McNamee, A. Haripriyan, D. Castillo, S. Grandy, A. Kessi, Measuring and modeling microclimate impacts of Sequoiadendron giganteum, Sustain. Cities Soc. 38 (2018) 509–525.

1055 https://doi.org/10.1016/j.scs.2017.12.028.
1056 [60] S.H. Lee, S.U. Park, A vegetated urban canopy model for meteorological and
1057 environmental modelling, Boundary-Layer Meteorol. 126 (2008) 73–102.
1058 https://doi.org/10.1007/s10546-007-9221-6.
1059 [61] H. Simon, J. Lindén, D. Hoffmann, P. Braun, M. Bruse, J. Esper, Modeling
1060 transpiration and leaf temperature of urban trees – A case study evaluating the
1061 microclimate model ENVI-met against measurement data, Landsc. Urban Plan.
1062 174 (2018) 33–40. https://doi.org/10.1016/j.landurbplan.2018.03.003.
1063 [62] J. Deng, B.J. Pickles, A. Kavakopoulos, T. Blanusa, C.H. Halios, S.T. Smith, L.
1064 Shao, Concept and methodology of characterising infrared radiative
1065 performance of urban trees using tree crown spectroscopy, Build. Environ. 157
1066 (2019) 380–390. https://doi.org/10.1016/j.buildenv.2019.04.056.
1067 [63] British Trees: native and non-native trees.
1068 https://www.woodlandtrust.org.uk/visiting-woods/trees-woods-and-
1069 wildlife/british-trees/ (accessed June 21, 2019).
1070 [64] C.F. Dietrich, Uncertainty, calibration and probability: The statistics of scientific
1071 and industrial measurement, 2nd ed., New York, 1991.
1072 https://doi.org/10.1201/9780203734759.
1073 [65] R Core Team, R: A language and environment for statistical computing, (2019).
1074 https://www.r-project.org/.
1075 [66] W.H. Oksanen J, Blanchet FG, Friendly M, et al., vegan: Community Ecology
1076 Package. R package version 2.5-6, (2019). https://cran.r-

project.org/package=vegan.

[67] M.F. Kassambara A, factoextra: Extract and Visualize the Results of Multivariate Data Analyses, (2017). https://cran.r-project.org/package=factoextra.

[68] S. Lê, J. Josse, F. Husson, FactoMineR: An R package for multivariate analysis, J. Stat. Softw. 25 (2008) 1–18. https://doi.org/10.18637/jss.v025.i01.

[69] P. Legendre, L. Legendre, Numerical Ecology, 3rd Editio, Elsevier, Amsterdam, Netherlands, 2012. https://www.elsevier.com/books/numerical-ecology/legendre/978-0-444-53868-0.

[70] J. Cavender-Bares, J.E. Meireles, J.J. Couture, *et al.*, Associations of leaf spectra with genetic and phylogenetic variation in oaks: Prospects for remote detection of biodiversity, Remote Sens. 8 (2016). https://doi.org/10.3390/rs8030221.

[71] ImageJ User Guide. https://imagej.nih.gov/ij/docs/guide/index.html (accessed September 20, 2019).

[72] Reading, ENG, United Kingdom — Sunrise, Sunset, and Moon Times. https://www.timeanddate.com/astronomy/uk/reading (accessed June 10, 2019).

[73] Horse chestnut leaf-mining moth. https://www.rhs.org.uk/advice/profile?pid=533 (accessed September 8, 2019).

[74] D.S. Falster, M. Westoby, Leaf size and angle vary widely across species: What consequences for light interception?, New Phytol. 158 (2003) 509–525. https://doi.org/10.1046/j.1469-8137.2003.00765.x.